

Author's final ACCEPTED VERSION

ACCEPTED IBIS 25<sup>TH</sup> June 2017

**Low and annually variable migratory connectivity in a long distance migrant:  
Whinchats *Saxicola rubetra* may show a bet-hedging strategy**

EMMA BLACKBURN<sup>1,2</sup>, MALCOLM BURGESS<sup>3</sup>, BENEDICTUS FREEMAN<sup>2,6</sup>, ALICE RISELY<sup>4</sup>, ARIN  
IZANG<sup>2</sup>, SAM IVANDE<sup>1,2</sup>, CHRIS HEWSON<sup>5</sup> & WILL CRESSWELL<sup>1\*,2</sup>

<sup>1</sup>*Centre for Biological Diversity, University of St Andrews, Harold Mitchell Building, St Andrews, Fife  
KY16 9TH, UK*

<sup>2</sup>*AP Leventis Ornithological Research Institute, Jos, Nigeria*

<sup>3</sup>*RSPB Centre for Conservation Science, The Lodge, Sandy, Beds SG19 2DL, UK*

<sup>4</sup>*Centre for Integrative Ecology, Deakin University, Geelong, VIC 3216, Australia*

<sup>5</sup>*British Trust for Ornithology, The Nunnery, Thetford, IP24 2PU UK*

<sup>6</sup>*Fauna & Flora International (FFI), 2 Harmon's Compound, Oldest Congo Town, Monrovia, Liberia*

\*Correspondence author: wrlc@st-and.ac.uk

**Keywords:** climate change, long-distance migrant, migrant declines, migratory spread, Palearctic,  
serial residency hypothesis.

Running head: Low and variable connectivity in Whinchats

The spatial scale of non-breeding areas used by long distance migrant animals can vary from specific, relatively small non-breeding areas for each independent breeding population (high connectivity) to a distribution over a large non-breeding area with mixing of breeding populations (low connectivity). Measuring variation in the degree of connectivity and how it arises is crucial to predict how migratory animals can respond to global habitat and climate change because low connectivity is likely an adaptation to environmental uncertainty. Here, we test whether use of non-breeding areas in a long distance migrant may be stochastic by measuring the degree of connectivity, and whether it is annually variable. 29 wintering Whinchats tagged with geolocators over two years within 40 km<sup>2</sup> in central Nigeria were found breeding over 2.549 million km<sup>2</sup> (26% of the land area of Europe), without an asymptote being approached in the relationship between area and sample size. Ranges differed in size between years by 1.506 million km<sup>2</sup> and only 15% of the total breeding range across both years overlapped (8% overlap between years when only first year birds were considered), well above the range size difference and below the proportion of overlap that would be predicted from two equivalent groups breeding at random locations within the observed range. Mean distance between breeding locations (i.e. migratory spread) differed significantly between years (2013, 604  $\pm$  18 km; 2014, 869  $\pm$  33 km). The results showed very low and variable connectivity that was reasonably robust to the errors and assumptions inherent in the use of geolocators, but with the caveat of only two years' ranges to compare, and the sensitivity of range to the breeding locations of a small number of individuals. However, if representative, the results suggest the scope for between-year variation (cohort effects) to determine migrant distribution on a large scale. Furthermore, for species with similar low connectivity, we would predict breeding population trends to reflect average conditions across large non-breeding areas: thus, as large areas of Africa become subject to habitat loss, migrant populations throughout Europe will decline.

Understanding how breeding and non-breeding sites are connected (migratory connectivity) is crucial for successfully predicting the response of migratory populations to environmental change (Bauer et al. 2016, Webster et al. 2002), particularly in the light of the widespread decline of migratory animals and increasing habitat loss and degradation (Gilroy et al. 2016, Flockhart et al. 2015, Costa et al. 2012, Vickery et al. 2014). Indeed, strategies to address migratory declines are often based on the idea that migrant populations will have connected breeding and non-breeding (or wintering) ranges and that the identification of these ranges is a priority (Runge et al. 2014, Runge et al. 2015, Martin et al. 2007). However, the scale of connectivity within and between populations of migrants and how this scale arises is still poorly known (Bauer et al. 2016).

The spatial scale of use of non-breeding areas by long distance migrant animals can vary from high connectivity - specific non-breeding areas for each independent breeding population – to low connectivity - a spread over a large non-breeding area and so mixing across breeding populations (Webster et al. 2002, Newton 2008). High connectivity is thought to arise from deterministic, targeted migration which allows population specialisation, but which reduces the resilience of a breeding population to changes in the distribution of non-breeding habitat, and so should only be selected for when the location and suitability of non-breeding sites are highly predictable and stable (Reilly & Reilly 2009, Cresswell 2014). In contrast, low connectivity is thought to arise from bet-hedging - untargeted migration which allows populations to deal with unpredictable environmental conditions (Reilly & Reilly 2009, Cresswell 2014, see also Botero et al. 2015). Breeding populations which are distributed over a large area during the non-breeding season should have greater potential to encounter suitable habitat over a larger range, even as this shifts due to climate change (Cresswell 2014). Furthermore, breeding populations that mix during the non-breeding season are likely to respond synchronously to changing conditions outside of the breeding season, whereas highly connected populations will respond independently (Esler 2000).

Long-distance migration of birds between northern Europe and sub-Saharan Africa is a recently evolved adaptive response to dynamic global climatic conditions (Cresswell et al. 2011, Fryxell & Holt 2013), suggesting that a bet-hedging strategy of individuals from the same brood migrating to geographically separate non-breeding sites might actually be the norm because of the resilience this strategy grants against dynamic conditions (Botero et al. 2015, Reilly & Reilly 2009). Consequently, migrants are most likely selected to be generalists within their wintering habitat (Ivande & Cresswell

2016), with low and variable migratory connectivity (the serial residency hypothesis - Cresswell 2014). A recent analysis of all migratory tracks that were available from tagged birds from 45 ecologically diverse land-bird species (including Passeriformes, Apodiformes, Falconiformes, Cuculiformes and Coraciiformes) in both Europe-Africa and the Americas flyways showed that connectivity at a large scale was the strategy used by most species (Finch et al. 2017). Here, we test whether use of non-breeding areas in a long distance migrant is likely to arise from a bet-hedging strategy by measuring degree of connectivity and whether it is annually variable.

Testing theories of migratory connectivity on a sufficiently large scale has recently become possible because of advances in tracking technology such as satellite and GPS tags: geolocators (light-level loggers) now allow animals of relatively small body size to be tracked on global scales for the first time (Bridge et al. 2013). We measured and tested migratory connectivity in the Whinchat *Saxicola rubetra*, a common but declining Afro-Palearctic long distance migrant bird (Müller et al. 2005, Britschgi et al. 2006). Uniquely, we measured connectivity: (i) from a location on the wintering ground to the breeding grounds, which allows much greater precision of locations (because of more variable daylength at higher latitudes, and because birds are highly likely to be stationary during the breeding period, Lisovski et al. 2012), (ii) in two separate years, (iii) with relatively large sample sizes (for geocator studies, Bridge et al. 2013), (iv) in a system where we know there are no effects of tagging (Blackburn et al. 2016), and (v) where there is very high non-breeding site fidelity (Blackburn & Cresswell 2016b). We tested the following two predictions that arise directly if non-breeding site selection has an important bet-hedging or stochastic component for first year birds, followed by site fidelity for adults to the sites which allowed their non-breeding survival (Cresswell 2014):

1. There will be an extensive range (on the scale of >1000km) of breeding locations sampled from individuals sampled at a single wintering location (Fig. 1).

2. There will be differences in breeding ranges between years when comparing samples from the same wintering location (Fig. 2). This may be particularly pronounced when comparing cohorts (i.e. first year birds only) because, in each year we would expect birds from spatially closer breeding populations to be affected in a similar way by conditions on migration (and so tending to end up in similar non-breeding areas), but for these conditions to vary between years (particularly because of large-scale annually variable timing of breeding or possibly variation in departure post-breeding) so that these non-breeding areas might shift in location annually.

We tested these predictions by mapping annual ranges and comparing how these differed in terms of size, overlap and mean distance between individuals, with the expectation that annual ranges will be large but variable between years.

## **METHODS**

The study took place between February 2013 and November 2013 (Year 1) and February 2014 until April 2015 (Year 2) during the dry season (early September to late April) on the Jos Plateau in the guinea savannah zone of central Nigeria, West Africa (N09°53', E08°59', approximately 1250 m altitude). Some colour-ringed only Whinchats were captured outside of these months (i.e. earlier in the wintering period or were colour-ringed birds that had returned from previous winters) to evaluate whether the geolocators affected survival (see Blackburn et al. 2016). Whinchats were captured within an area of approximately 5 x 8 km; full site details are described in Blackburn and Cresswell (2016b). Capture areas were principally open scrubland with varying degrees of habitat degradation from human habitation, arable farming and livestock grazing, the latter increasing in intensity over the dry season (see Hulme & Cresswell 2012, Blackburn & Cresswell 2015). The study area represents typical wintering habitat for this species in the region (open savannah) and has high densities of Whinchats.

Whinchats were caught with spring traps and mist nets in late February and March in 2013 or 2014 (birds in 2014 were on average tagged  $10.3 \pm 3.3$  SE days earlier than in 2013,  $F_{1,27} = -3.2$ ,  $P = 0.004$ ). Birds were aged and sexed (Jenni & Winkler 1994), ringed with unique combinations of colour-rings, and fitted with a geocator. In Year 1, we deployed 49 and in Year 2 we deployed 131 geolocators fitted using leg-loop 'Rappole-Tipton' (also called backpack) harnesses. Full details of tag and harness design are given in Blackburn et al. (2016), but importantly for this study, there was variation between years in the length of light stalks of the tags. In 2013 all tags had a 10mm length light stalk, whereas in 2014 tags had either no light stalks, 5mm or 10mm length light stalk (see below). Tags weighed on average 0.63 g (0.01SE), representing 4.1 % of average body mass. There was no overall significant reduction in between-year resighting rate (our proxy for survival, Blackburn & Cresswell 2016b) comparing tagged and untagged birds in either year (Blackburn et al. 2016).

Attempts were made to recapture any returning tagged bird resighted in the following winter. Upon recapture, geolocators were removed by cutting the harness and birds were released unharmed after briefly assessing body condition (see Blackburn et al. 2016). Sample sizes of breeding locations used

in this paper are less than the number of individuals that returned with geolocators because 18/39 returning birds in 2014 could not be recaptured (because many had become extremely wary of spring-traps and mist-nets), two individuals had lost their loggers because of harness failure, and several loggers in 2013 suffered battery failure before the birds reached the breeding ground. Overall we include all possible data from 29 individuals (12 in 2013 and 17 in 2014 – any individual was tracked only in one year); data came from eight females and 21 males (ratio was not significantly different by year, Fisher's exact test  $P = 0.41$ ), and 13 adults (see Supplementary Material Figure S2 for the locations of these individuals), 13 first years and three that could not be aged confidently (ratio of known age birds was not significantly different by year, Fisher's exact test  $P = 0.43$ ). In 2014, of the recaptured birds, 4 had tags with no light stalk (see Supplementary Material Figure S2 for the locations of these individuals), 8 with 5 mm light stalks and 5 with 10 mm light stalks.

## **Analyses**

Raw data were downloaded, viewed and preliminarily cleaned using the BASTrack software suite (British Antarctic Survey, Cambridge, UK; see Fox 2010 for an overview of the following processes). We adjusted for clock drift, assuming that any drift was linear. We used the Transedit2 software that is part of the BASTrack software to view raw data as light curves over time. We used a threshold value of 2 to define sunrise, which is close to civil twilight. False twilight events due to shading from weather or vegetation were identified and removed with the 'minimum dark period' filter (we used 4 hours), which removes any impossible sunrise and sunset events (for a review and exploration of the effects of environmental factors on geolocator data, see Lisovski et al. 2012). Data were then visually inspected to ensure that only one sunrise and sunset occurred within any 24-hour period.

Further analyses were then carried out using R 2.13.1 (R Development Core Team 2014). We used the LoessFilter in the R package 'GeoLight' (Lisovski & Hahn 2012) to validate twilight events and identify those influenced by shading events at dawn and dusk: a polynomial regression of twilight events identified residuals that were greater than three times the interquartile range and are therefore likely to be erroneous (Lisovski & Hahn 2012). Identified outliers were checked within the original data and were retained because large movements resulting from migration were sometimes incorrectly identified as outliers, and outliers were reasonably normally distributed in the winter and summer periods analysed (and so their effects cancelled out when calculating mean locations).

Because conditions away from the wintering grounds are unknown, we carried out two calibrations to determine the correct sun elevation angle value (henceforth 'SEA', which influences the determined latitude) for winter (SEAw) and summer (SEAs). For SEAw, we used the LocatorAid software from the BASTrack software suite, which uses known residency times and wintering location to calculate the corresponding wintering ground SEA value. We attempted to find the correct summer angle using the Hill-Ekstrom (H-E) calibration method (Hill & Braun 2001, Ekstrom 2004, see Tottrup et al. 2012) in which we selected the SEA value that gave the least amount of variation in latitude. If the calibration was not successful, we used the mean value for all of the loggers for which the calibration had been successful, calculated for each year separately (see Supplementary Material Figure S1 for a mapped description of which birds were affected; 4 in 2013 and 12 in 2014, and the locations associated with each SEA value used). Because the mean SEA value may have differed between years (e.g. because of inter-annual differences in cloud cover), we also simply used the same average SEA value for all birds regardless of year in a further analysis but with little change to the results (see below). In further sensitivity analyses (see below), we also varied the SEA value for these 'mean value' birds over the range of observed values from the Hill-Ekstrom calibration for that year to demonstrate that this uncertainty (and the assumption of using mean values for those birds where the calibration did not give clear results) made little difference to the results (see below). Once a corresponding SEAw and SEAs value was known for each logger, noon and midnight locations for a two week period when variance in locations was minimal (and all Whinchats should have been stationary on their breeding grounds in any case) were derived from sunrise and sunset times using the 'coord' function in the Geolight package (Lisovski & Hahn 2012). Breeding location was further confirmed from visual inspection of latitude and longitude changes with date and through identification of stationary periods using Geolight. We used the mean of all summer locations (calculated using SEAs values) for the first two weeks in June for all birds except two which showed evidence of some small movement in June where we used the last two weeks in May, and one bird that showed evidence of major movement in June, to a new stable location in July (perhaps after breeding failure) where we used the first two weeks in July ('Best' Analysis 1). All two week 'most' stationary locations identified were at very similar locations to the mean of all locations from mid-May to mid-July for all but the three birds which moved during this period, and for these three birds, the 'second' location was then stable for several weeks afterwards probably indicating breeding. Using these 'best' most stationary periods, or simply all locations from each bird in June gives very similar results for all analyses (See Supplementary

Material Tables S1 and S2, compare results for Analysis 1 & 4); other periods between mid-May to the end of July gave similar results (unpubl. data).

Maps were plotted using the raster, rgdal, rgeos and mapproj libraries in R. Minimum convex polygon areas, overlaps of these polygons and distances (considering the curvature of the earth) were calculated using the alphahull and geosphere libraries in R. Areas were calculated for both years combined and for 2013 and 2014 separately; the range overlap between the two years as a percentage of the overall range (PHR) was also calculated. More sophisticated analyses using kernels generated from the density of points were not appropriate because we had only a single breeding location for each bird (albeit with variation due to geolocator errors in precision and SEA assumptions) and so sample sizes of overall points split by year were too small for meaningful calculations of probability density functions.

Ranges are sensitive to sample size with range increasing with sample size until an asymptote is reached when individuals from nearly all parts of the range are part of the sample. We tested the degree to which the ranges which we measured approached an asymptote, i.e. whether the range calculated from our small samples likely approached the true range. The effect of sample size on the range recorded overall and in each year separately was tested by randomly sub-sampling the breeding locations for all birds, selecting 4 points up to  $N - 1$  points overall and for the individual years, 1000 times and then calculating the range for each sub-sample. The relationship between mean range of the 1000 sub-samples and number of tagged birds was then tested with a GLM to determine the line of best fit, comparing linear and quadratic fits to determine if an asymptote in range with sample size had been reached.

We tested whether there was any significant difference in the size of the range and degree of overlap of range between the two years by creating comparison data sets of the range size and overlap that would occur if two groups of random points were selected from the total breeding range observed in the two years. (i) 29 geographic points were selected randomly from the combined observed range plotted for all individuals across both years; the first 12 points were placed in group 1 corresponding to the first year's sample and the final 17 points in group 2 corresponding to the second year's sample. Minimum convex polygons for each group and the combined groups were calculated using the alphahull library in R; the area in each year and the difference in area between years, overall area and the overlap in area between the two groups as a percentage of this total area for both groups was



then calculated. The procedure was repeated 1000 times to determine the 95% confidence intervals of area difference and overlap for randomly selected points (i.e. with random difference in location between years). (ii) We repeated (i) with 13 points randomly selected (7 in group 1 and 6 in group 2) to create a similar random comparison data set considering only the first year birds.

We also measured the mean distance between breeding locations in each year (equivalent to migratory spread as calculated in Finch et al. 2017). A matrix of all possible pairs of distances was created and the distance between locations was calculated using the `distHaversine` function in R, and then the mean and standard error of these annual mean distances was calculated. We tested whether there was any significant difference in mean spread of locations between years using a t test with degrees of freedom conservatively set to the number of original locations (i.e.  $N = 29$  for all birds and  $N = 13$  for first year birds).

We tested for any systematic biases in latitude or longitude calculated with respect to variation in sex, age, stalk length, SEA value, and whether this SEA value was estimated using the H-E calibration or an average, in a GLM controlling for year. We also included interactions with year in this model to test whether any effects of these potential biases acted differently in the two years of the study. There were no significant effects (systematic biases that may have accounted for the results) in any variables or interactions apart from light stalk length (See Supplementary Material Table S3 & S4). A significant effect of light stalk length on latitude (but not longitude) was found so that more southerly latitudes were recorded from tags without light stalks, dependent on the SEA value considered (details in Table 2). Correspondingly we adjusted all latitudes for the light stalk effect so that if a tag did not have a light stalk we added the parameter estimate for light stalk (between +3.8 to +5.0 degrees of latitude – see Table 2) to the estimate of latitude: note this had the effect of reducing any differences between years, because 2014 locations were generally to the south of 2013 locations (see Figures in Supplementary Material and particularly Figure S2 which identifies those loggers that had no light stalks). Results using stalk length as a three-way factor, or as a two-way factor of light stalk presence or absence, were similar, and models which considered stalk length as a three-way factor were more than 2 AIC units worse than identical models with stalk length as a two-way factor. We therefore only considered the presence and absence of a light stalk when considering the potential confounding effects of tag design in detail. There was no significant variation in latitude with age or sex in any model and models were substantially ( $\Delta AIC > 2$ ) improved by the removal of these terms

which allowed the full data set to be used (i.e. by including those birds that lacked age data; Supplementary Material Table S3 & S4). Final adjustment of latitude was therefore calculated from a model including year and light stalk presence or absence (Supplementary Material Table S3 & S4). The appropriate adjustment (i.e. depending on SEA value) was used for each analysis and its associated range of SEAs values (Table 2). Because of the potential confounding effect of light stalk between the years that we identified, all analyses were carried out on both the raw observed location data set (Table 1) and also the data set with locations adjusted for the absence of a light stalk (Table 2).

Because of uncertainty introduced by not knowing breeding location SEA value, we also analysed the data under a set of scenarios with different assumptions for the SEA value (Supplementary Material Tables S1 & S2). These analyses were: 1. 'Best' with summer SEAs values determined using the Hill-Ekstrom calibration, or the mean value from the calibrations for that year for birds where the calibration was not conclusive, for the 2 week breeding season period with lowest variation in latitude; 2. Winter SEAw value for the 2 week breeding season period with the lowest variation in latitude; 3. Average SEA value - the mean location calculated for each bird for sun elevation angles 2 to 6 at 0.5 increments (i.e. 9 mean locations), then averaged across these 9 locations, applied to all birds regardless of year; 4. As analysis 1, but for *all* locations in June; 5. As analysis 1, but with the maximum or minimum SEA value observed for that year used for those birds where the Hill-Ekstrom calibration was not conclusive, so that the range for each year was minimised (i.e. an analysis to examine the minimum range possible under the uncertainty of the SEAs values); 6. As analysis 1, but with the maximum or minimum SEA value observed for that year used for those birds where the Hill-Ekstrom calibration was not conclusive, so that overlap for each year was maximised (i.e. an analysis to examine the maximum overlap possible under the uncertainty of the SEAs values).

In short, extensive sensitivity analyses (Tables S1 & S2 and illustrated in Figures S3 – S20 in Supplementary Material) - regardless of SEA value and light stalk adjustment assumptions – showed very similar results. The results presented here use the data set which we considered to be the most accurate (Analysis 1. 'best' data set), but these results are representative of other scenarios which vary the assumptions of sun elevation angle; these are also detailed in Supplementary Material Tables S1 & S2.

Best models were identified on the basis of AIC, and models differing in AIC of less than 2 were considered as equally valid (Burnham and Anderson 2002). Model fits were evaluated from diagnostic model plots, and models were presented if assumptions were reasonably met (Crawley 2007). Mean values are presented with one standard error (SE) in all cases;  $R^2$  values were adjusted in all cases.

## Results

### *Overall range extent*

Whinchats tagged within 40 km<sup>2</sup> of central Nigeria had a breeding range extent of up to 2.549 million km<sup>2</sup> (range 2.351 – 2.549 million km<sup>2</sup>; note here and subsequently, range is calculated across the different SEA values, see Supplementary Material Tables S1 & S2), or 26% of the land area of Europe (Table 1, Figure 3). The latitudinal range of the breeding grounds was approximately 16.3 degrees from Serbia to the southern border of Finland (a north to south distance of approximately 1,700 km) and the longitudinal range was 35.9 degrees from Poland to the Ural Mountains in Russia (an east to west distance of approximately 2,300 km; Fig. 3). When values were adjusted for the effect of light stalk absence, breeding range extent was 1.681 million km<sup>2</sup> (1.681 – 1.835 million km<sup>2</sup>), or 17% of the land area of Europe (Table 2).

The total breeding range observed overall did not reach an asymptote with sample size. A straight line relationship between range and sample size for all birds unadjusted for light stalk absence gave a reasonable model fit (Fig. 4a). When latitudes were adjusted for the absence of light stalks, a quadratic fit was much better for the relationship between range and sample size, but although the relationship was becoming less steep it did not approach an asymptote (Fig. 4b).

### *Annual variation in range extent*

Whinchats had larger ranges in 2014 compared to 2013. The breeding area of all Whinchats tagged in central Nigeria, unadjusted for light stalk absence, was much lower in 2013 (0.651 million km<sup>2</sup>, range 0.503 – 0.744) than that of the range in 2014 (2.157 million km<sup>2</sup>, range 1.971 – 2.230; Fig. 5). The differences in unadjusted area between the two years were greater than expected by chance for all SEA analyses for all birds (Table 1, Supplementary Material Table S1). The breeding area of all Whinchats tagged in central Nigeria in 2013 was much smaller than that of the range adjusted for light stalk absence in 2014 (1.650 million km<sup>2</sup>, range 1.650 – 1.814); the differences in adjusted area between the two years were greater than expected by chance for all SEA analyses (Table 2,

Supplementary Material Table S2). For first year birds only, the breeding area unadjusted for light stalk absence in 2013 (0.470 million km<sup>2</sup>, range 0.429 – 0.518) was much smaller than that of the range in 2014 (1.199 million km<sup>2</sup>, range 1.145 – 1.199; Fig. 3). The differences in area between the two years were greater than expected by chance for all SEA analyses of first year birds (Table 1, Supplementary Material Table S1). For first year birds only, adjusted for light stalk presence in 2014, the breeding area in 2013 was much smaller than that of the 2014 area (1.232 million km<sup>2</sup>, range 0.829 – 1.232). The differences in area between the two years were greater than expected by chance for all SEA analyses of first year birds, apart from when using the summer SEA for all locations in June (Table 2, Supplementary Material Table S2).

Whinchat ranges in the two years generally overlapped relatively little, although this was less clear in first year birds. The breeding range of all Whinchats showed relatively little overlap between years: only 15% (range 8.7% - 27.1%) of the total range for both years was overlap, with all overlaps being lower than that expected by chance (Table 1, Supplementary Material Table S1). For latitudes adjusted for light stalk absence, overlap was 37% (range 31% - 42%) with a trend ( $P = 0.03$  to  $P = 0.07$ ) for overlaps to be lower than that expected by chance (Table 2, Supplementary Material Table S2). For first year birds only, there was also a trend for relatively low overlap in the breeding ranges between years: 8% (range 2.7% - 18.9%) of the total range for both years overlapped, with overlaps only being lower than that expected by chance when using the summer SEA for all locations in June, and marginally significant for the 'best' analysis (Table 1, Supplementary Material Table S1). For latitudes adjusted for light stalk absence for first year birds, however, there was reasonable overlap in the breeding ranges between years: 33% (range 3.3% - 39%) of the total range for both years overlapped, with all overlaps only being lower than that expected by chance when using the summer SEA for all locations in June (Table 2, Supplementary Material Table S2).

The mean distance between breeding locations was greater in 2014. For all birds, the mean distance was 604 km in 2013 (range 591 – 631) and was 869 km in 2014 (range 839 – 903) and the difference was highly significant for all SEA analyses (Table 1, Supplementary Material Table S1). For all birds, the mean distance adjusted for light stalk in 2014 was 820 km (range 817 – 828) and the difference was highly significant for all SEA analyses (Table 2, Supplementary Material Table S2). For first year birds, the mean distance was 598 km in 2013 (range 567 – 631) and was 1059 km in 2014 (range 1013 – 1060) and the difference was highly significant for all SEA analyses (Table 1, Supplementary

Material Table S1). The mean distance adjusted for light stalk in 2014, for first year birds, was 1042 km (range 950 – 1042) and the difference was significant for all SEA analyses (Table 2, Supplementary Material Table S2).

The total breeding range observed in each year separately did not reach an asymptote with sample size. For all birds, the relationship between range and sample size was dependent on sample size, but with different slopes dependent on year (interaction of year\* sample size  $t_{1,19} = 11.5$ ,  $P < 0.0001$ ). There was therefore a significantly greater spread in breeding locations for 2014 ( $0.141 \pm 0.007$ ) compared to 2013 ( $0.061 \pm 0.006$ ), i.e. a much higher rate of increase in range per increase in sample size (Fig. 4c). The relationships between range and sample size were linear (overall model  $R^2 = 0.99$ ). The same result was found using 2014 range adjusted for the effect of light stalk absence. For all birds adjusted for the effect of light stalk, the relationship between range and sample size was dependent on sample size, but with different slopes dependent on year (interaction of year\*sample size  $t_{1,19} = 11.7$ ,  $P < 0.0001$ ): there was therefore a significantly greater spread in breeding locations for 2014 ( $0.109 \pm 0.004$ ) compared to 2013 ( $0.061 \pm 0.004$ ): i.e. a much higher rate of increase in range per increase in sample size (Fig. 4d). The relationships between range and sample size were linear (overall model  $R^2 = 0.99$ ).

## DISCUSSION

We made two predictions: a large range and an annually variable range. Both were supported by our results. Our results probably show very low connectivity, i.e. connectivity only at a very large scale: Whinchats from a single wintering location were spread out over a continental scale in Europe (approximately 40% of the maximum east to west width of Europe). The reverse, that individuals from the same breeding location in Europe will be spread out over a large non-breeding area in Africa, is logically very likely to apply (Fig. 1), especially in a species with such high non-breeding site fidelity (Blackburn & Cresswell 2016b), breeding ground site fidelity (Bastian 1992) and a non-breeding distribution across the whole of Africa, from Senegal to Tanzania. Whinchats from any local breeding population in the east of Europe are therefore likely to spread out over a similar continental scale in Africa (33% of the maximum east to west width of Africa at typical wintering latitude, 8 degrees). Furthermore, we suggest that the distribution of a wintering population of Whinchats in Africa may also be annually variable, with potential shifts in average occurrence between years of the order of hundreds of kilometres.

Our results, at least in terms of the large scale of the ranges and variation we have described, seem reasonable. Although locations from geolocators have errors, these errors within individual stationary periods were reasonably normally distributed, as probably were errors between individuals through incorrectly assessing SEA value (e.g. some individuals will have lower and some higher SEAs values than they should, but these will likely be random with respect to individual location on a large scale). There were, for example, no correlations between SEA value used in the 'best' analysis and latitude or longitude (model of SEAs ~ latitude + longitude; latitude  $0.04 \pm 0.03$ ,  $t = 1.2$ ,  $P = 0.24$ , longitude  $-0.002 \pm 0.014$ ,  $t = -0.2$ ,  $P = 0.86$ ;  $F_{2,26} = 0.7$ ,  $R^2 = 0$ ). Furthermore, our analyses (analyses 2 & 3, see Supplementary Material Tables S1 & S2) that simply used an average value of SEA for all birds will have increased the randomness of the errors, and both gave very similar results to analyses which used the best available information to more accurately estimate SEA value. Consequently, range areas and range differences between years should be real, even if imprecise in location. Fundamentally, the scale of geocator errors (Fudickar et al. 2012) is also relatively small compared to the scale of the ranges identified in this study. Our analyses, which used different assumptions regarding sun elevation angle values, breeding stationary period and effect of absence of light stalks, showed broadly similar results at the large scale of this study: despite variation in the individual breeding locations in each analysis, 2013 birds had a smaller spread and range than 2014 birds and there was reasonable evidence for a difference in range between years when comparing overlap.

Although the H-E calibration should have accounted for differences in shading between birds, locations and years, the effect of light stalk presence / absence on latitude demonstrated that this calibration was biased by shading (in this case, most likely through from flight feathers covering the light sensor). Consequently, the latitudes of breeding locations in each year could have been biased by the degree & variability of weather shading in the period for which data was used in the calibration, especially when this was based on a short period (2 weeks). This could potentially have caused spurious shifts in range between years. The difference in SEA between years after the H-E calibration was small (2014,  $-0.55 \pm 0.58$ ,  $t_{1,9} = -1.0$ ,  $P = 0.37$ ) and would have shifted 2014 locations by about 0.6 degrees latitude to the south away from 2013 locations, thus the bias, if it was operating, would have increased any differences in range and overlap between years for the 'best' analysis (Analysis 1). It could not, however, completely explain the differences in spread of individuals between years because this also has a longitudinal component (albeit due to a few birds). In any case, any such bias was completely removed when using the same average SEA value for all birds regardless of year.

We still found variation when controlling for cohort effects, despite smaller sample sizes. The larger population considered contained both first year and adult birds from previous years, each individual year having a likely different cohort range effect. Any comparison between two consecutive years that includes adult birds will therefore be conservative because, as more cohorts are included, the range in any one year will regress to the mean range of the population. This will act to blur any differences between years. However, Whinchat annual survival rate for this population is of the order of 54% (Blackburn & Cresswell 2016a), so even when considering adults, it was likely that over 75% of the population sampled in each year consisted of only two cohorts. Unfortunately, any convincing analyses of only adult birds – which should show less variation (this was the case, but not included here) - was precluded by only four adults being part of the 2013 sample versus nine in the 2014 sample (see Supplementary Material Figure S2), so making any confirmation of the null hypothesis likely simply because of low power, regardless of any actual differences or not between years.

Although the results of the analyses are consistent regardless of assumptions, they rely on small sample sizes. Any range or migratory spread calculation will be dependent on its relatively few peripheral points, and the sample size for inter-annual comparisons is only two years. The overlap result is also difficult to interpret, being confounded by variation in size as well as a possible shift in location, although the study shows clear evidence of annual variation in measures that will affect overlap (i.e. migratory spread, and the relationship between range and sample size). The study clearly needs to be repeated, ideally tracking only first-year Whinchats from other areas in Africa in consecutive years to determine inter-annual variability in range, with sufficient breeding locations obtained to analyse range using probability density functions. Nevertheless, our results suggest that low and variable connectivity can occur in Whinchats. Our range estimates did not approach an asymptote with a sample size of 29 tagged birds and there were differences in range between years sufficient that conclusions about where Whinchats wintering in central Nigeria may have bred, or their degree of connectivity, would have differed on an international scale, if the study had been only conducted in a single year, or with a different sample size.

Our results show migratory spread consistent with the pattern just emerging from other studies. Long distance migrant birds have high migratory spread regardless of phylogeny and flyway; the predicted average maximum distance between wintering individuals from the same breeding population, across 45 species, is of the order more than 3,000 km (Finch et al. 2017). It should be noted, however, that

all species lie on a continuum of connectivity and there are examples of particularly low connectivity species (e.g. Pallid Harriers *Circus macrourus* Terraube et al. 2012) and particularly high connectivity species (e.g. Nightingales *Luscinia megarhynchos* Hahn et al. 2013); Whinchats would appear to lie closer to the extreme end of low migratory connectivity. Inter-annual shifts in range are less well documented or tested, even in the few multi-year datasets, but Liechti et al. (2015) reported that Barn Swallows *Hirundo rustica* wintered 400km to the north in one year compared to the population mean for the previous year, and Swainson's Thrush *Catharus ustulatus* (Delmore and Irwin, 2014) and Semi-collared Flycatcher *Ficedula semitorquata* (Briedis et al., 2016) populations also probably show inter-annual variability in mean wintering location.

Although the low and variable connectivity probably demonstrated here could arise through a bet-hedging strategy because first year non-breeding site selection is to a degree stochastic, other mechanisms may have caused the pattern. As detailed in Figure 1, if migrants with wide breeding ranges in Europe have relatively small wintering ranges in Africa (e.g. Eleonora's Falcon *Falco eleonora* Gschweng et al. 2008, Kassara et al. 2012), effectively concentrating in a small non-breeding area, and they are then sampled in that small area, they will show apparently very low connectivity. This is unlikely to be the case for Whinchat, however, due to its very large non-breeding distribution in Africa. Variation in breeding success on a continental scale may also cause differences in breeding ranges when sampled in different years from Africa. For example, northerly breeding Whinchats in 2012 may have produced more first years than southerly Whinchats, and vice-versa in 2013. Annual variation in settlement patterns for first breeding, particularly if a species has low or environmentally variable natal site fidelity (Studds et al. 2008), could also lead to the patterns observed here.

With the caveat that alternative explanations are possible, the results of this study are consistent with the predictions of the serial residency hypothesis, where low and variable connectivity emerges from juveniles having deterministic migratory directions varying around a mean 'approximate' migration direction and then subsequent non-breeding locations further varying due to annually variable events (e.g. breeding phenology and weather patterns during migration), followed by high wintering site fidelity of surviving adults (Cresswell 2014). Such a system makes sense in terms of the evolution of migration, which is an adaptation to large scale climate change and consequent shifting of suitable habitat (Cresswell et al. 2011, Fryxell & Holt 2013). As required resources or environments shift in



location, at least some offspring from a brood will find survivable routes, thus a bet-hedging strategy – where some offspring migrate in slightly different directions and/or different times (Reilly & Reilly 2009) - is adaptive when the location of suitable habitat is unknown and unpredictable (Cresswell 2014). As location of suitable conditions becomes more predictable, this bet-hedging strategy becomes less adaptive because a proportion of offspring will always miss the target and arrive in unsuitable habitat. Although adjustments within the wintering range can be made by further small or even large scale migratory movements, there may be costs because migration itself can represent a period of high mortality (Sillett & Holmes 2002, Strandberg et al. 2010).

A bet-hedging strategy is, however, unlikely to be adaptive if the availability of habitat declines, because an increasing proportion of offspring (or the population) will miss the target. Thus the connectivity optimum for climate change (bet-hedging leading to a wide non-breeding distribution) is the opposite of the connectivity optimum for habitat loss (with specific site or area selection leading to a narrow non-breeding distribution): migratory species cannot adapt to both. This creates an adaptive conflict in the face of both shifting habitat due to climate change and disappearing habitat because of rapidly increasing human populations in Africa. The many migrant species that apparently make a first naïve bet-hedging migration (i.e. excluding those relatively few species that use social learning and that migrate in mixed age groups) will have a wide non-breeding distribution and an annually increasing proportion of their populations arriving in unsuitable habitat on the wintering ground, thus reducing survival. The wide non-breeding distribution ensures that *any* reduction in habitat availability over a very wide area of the wintering grounds affects *all* populations over a very wide breeding area (Sutherland & Dolman 1994, Taylor & Norris 2010, Cresswell 2014). This general characteristic may then significantly contribute to the decline of so many populations of migratory species (see Vickery et al. 2014), regardless of their location on the breeding grounds and their ecology.

A number of predictions arise from this framework with respect to population trends and distribution. First, population trends will be a function of a population's distribution or migratory spread on the non-breeding ground: at some point the proportion of first years that miss the target must reduce juvenile survival below the level of productivity/recruitment, although this will vary dependent on the degree of non-breeding habitat/resource specialisation (i.e. how big the target is, or how costly it is to miss it). Species with low migratory spread will have favourable population trends where they winter in climatically stable areas and are likely to be habitat specialists; selection may act to reduce genetic

variability in migration direction and/or to promote social learning of routes. Species with high migratory spread will have favourable population trends (e.g. see Gilroy et al. 2016), particularly if they winter in climatically variable areas and are likely to be habitat generalists (e.g. Hulme & Cresswell 2012, Blackburn & Cresswell 2015); selection may therefore act to increase variability in migration direction. Consequently, species with intermediate migratory spreads and generalist habitat requirements might have more favourable population trends. Second, appreciating that there are cohort effects will greatly affect our understanding of distribution. Range should shift rapidly on the wintering ground so that range maps built from sight records over decades may not accurately reflect current range, or indeed the actual range in any year. Furthermore, models to predict the habitat requirements of migrants at a large scale will be subject to errors both from the false negatives from stochastic range shifts as well as the 'false' positives of juvenile cohorts in unsuitable areas with much lower survival.

Testing such predictions is essential if we are to respond to the extremely rapid decline of migratory species across global flyways. Whether we adopt species-specific and local site specific land sparing solutions, rather than widespread land sharing solutions likely depends on degree of connectivity and how it affects survival.

The study was carried out in Nigeria where no licences are required for the procedures used. Nevertheless this study was carried out under the ethical guidelines of the AP Leventis Ornithological Research Institute Scientific Committee (APLORI is the only ornithological research institute in Nigeria) based on the Association for the Study of Animal Behaviour guidelines and those of the British Trust for Ornithology's ringing scheme. All personnel involved in fieldwork – either catching, colour-ringing or tagging birds had BTO ringing licences. MB had been previously licensed to fit geolocators in the UK. This work was supported by the Chris Goodwin, A.P. Leventis Conservation Foundation, AP Leventis Ornithological Research Institute, the British Ornithologists' Union and the Linnean Society. This is paper number (to be completed at proof stage) from the AP Leventis Ornithological Research Institute.

## REFERENCES

- Bastian, H.-V. 1992. Breeding and natal dispersal of Whinchats. *Ringling & Migration*, **13**: 13-19.
- Bauer, S., Lisovski, S. & Hahn, S. 2016. Timing is crucial for consequences of migratory connectivity. *Oikos*, **125**: 605-612.
- Blackburn, E., Burgess, M., Freeman, B., Risely, A., Izang, A., Ivande, S., Hewson, C. & Cresswell, W. 2016. An experimental evaluation of the effects of geolocator design and attachment method on between-year survival on Whinchats *Saxicola rubetra*, a long distance migrant. *Journal of Avian Biology*, **47**: 530-539.
- Blackburn, E. & Cresswell, W. 2015. Fine-scale habitat use during the non-breeding season suggests that winter habitat does not limit breeding populations of a declining long-distance Palearctic migrant. *Journal of Avian Biology*, **46**: 622-633.
- Blackburn, E. & Cresswell, W. 2016a. High overwinter and annual survival for a declining Palearctic migrant: evidence that wintering conditions may not limit migrant populations *Ibis*, **158**: 92-105.
- Blackburn, E. & Cresswell, W. 2016b. High winter site fidelity in a long-distance migrant: implications for wintering ecology and survival estimates. *Journal of Ornithology*, **157**: 93-108.
- Botero, C. A., Weissing, F. J., Wright, J. & Rubenstein, D. R. 2015. Evolutionary tipping points in the capacity to adapt to environmental change. *Proceedings of the National Academy of Sciences*, **112**: 184-189.
- Bridge, E. S., Kelly, J. F., Contina, A., Gabrielson, R. M., MacCurdy, R. B. & Winkler, D. W. 2013. Advances in tracking small migratory birds: a technical review of light-level geolocation. *Journal of Field Ornithology*, **84**: 121-137.
- Britschgi, A., Spaar, R. & Arlettaz, R. 2006. Impact of grassland farming intensification on the breeding ecology of an indicator insectivorous passerine, the Whinchat *Saxicola rubetra*: Lessons for overall Alpine meadowland management. *Biological Conservation*, **130**: 193-205.

548 Costa, D. P., Breed, G. A. & Robinson, P. W. 2012. New Insights into Pelagic Migrations: Implications  
 549 for Ecology and Conservation. In *Annual Review of Ecology, Evolution, and Systematics*, Vol  
 550 43. (ed. D. J. Futuyma), **Vol. 43**, pp. 73-96.

551 Crawley, M. J. (2007) *The R Book*, Chichester: John Wiley & Sons Ltd.

552 Cresswell, K. A., Satterthwaite, W. H. & Sword, G. A. 2011. Understanding the evolution of migration  
 553 through empirical examples. In *Animal migration: a synthesis*. (eds. E. J. Milner-Gulland, J. M.  
 554 Fryxell & A. R. E. Sinclair), pp. 1-16. Oxford, U.K.: Oxford University Press.

555 Cresswell, W. 2014. Migratory connectivity of Palaearctic-African migratory birds and their responses  
 556 to environmental change: the serial residency hypothesis. *Ibis*, **156**: 493-510.

557 Ekstrom, P. A. 2004. An advance in the geolocation by light. In *Memoirs of the National Institute of*  
 558 *Polar Research, Special Issue*. (ed. Y. Naito), pp. 210-226. Tokyo: National Institute of Polar  
 559 Research.

560 Esler, D. 2000. Applying metapopulation theory to conservation of migratory birds. *Conservation*  
 561 *Biology*, **14**: 366-372.

562 Finch, T., Butler, S., Franco, A. & Cresswell, W. 2017. Low migratory connectivity is common in long-  
 563 distance migrant birds. *Journal of Animal Ecology*: n/a-n/a.

564 Flockhart, D. T. T., Pichancourt, J. B., Norris, D. R. & Martin, T. G. 2015. Unravelling the annual cycle  
 565 in a migratory animal: breeding-season habitat loss drives population declines of monarch  
 566 butterflies. *Journal of Animal Ecology*, **84**: 155-165.

567 Fox, J. 2010. Geolocator manual v8 (March 2010).  
 568 [http://www.antarctica.ac.uk/engineering/geo\\_downloads/Geolocator\\_manual\\_v8.pdf](http://www.antarctica.ac.uk/engineering/geo_downloads/Geolocator_manual_v8.pdf).

569 Fryxell, J. M. & Holt, R. D. 2013. Environmental change and the evolution of migration. *Ecology*, **94**:  
 570 1274-1279.

571 Fudickar, A. M., Wikelski, M. & Partecke, J. 2012. Tracking migratory songbirds: accuracy of light-  
 572 level loggers (geolocators) in forest habitats. *Methods in Ecology and Evolution*, **3**: 47-52.

573 Gilroy, J. J., Gill, J. A., Butchart, S. H. M., Jones, V. R. & Franco, A. M. A. 2016. Migratory diversity  
574 predicts population declines in birds. *Ecology Letters*, **19**: 308-317.

575 Greenwood, P. J. & Harvey, P. H. 1982. The natal and breeding dispersal of birds. *Annual Review of*  
576 *Ecology and Systematics*, **13**: 1-21.

577 Gschweng, M., Kalko, E. K. V., Querner, U., Fiedler, W. & Berthold, P. 2008. All across Africa: highly  
578 individual migration routes of Eleonora's falcon. *Proceedings of the Royal Society B-Biological*  
579 *Sciences*, **275**: 2887-2896.

580 Hahn, S., Amrhein, V., Zehndt, P. & Liechti, F. 2013. Strong migratory connectivity and seasonally  
581 shifting isotopic niches in geographically separated populations of a long-distance migrating  
582 songbird. *Oecologia*, **173**: 1217-1225.

583 Hill, R. D. & Braun, M. J. 2001. Geolocation by light level - the next step: Latitude. In *Electronic*  
584 *tagging and tracking in marine fisheries*. (eds. J. R. Silbert & J. L. Nielsen), pp. 315-330.  
585 Dordrecht: Kluwer Academic publishers.

586 Hulme, M. F. & Cresswell, W. 2012. Density and behaviour of Whinchats *Saxicola rubetra* on African  
587 farmland suggest that winter habitat conditions do not limit European breeding populations.  
588 *Ibis*, **154**: 680-692.

589 Ivande, S. T. & Cresswell, W. 2016. Temperate migrants and resident bird species in Afro-tropical  
590 savannahs show similar levels of ecological generalism. *Ibis*, **158**: 496-505.

591 Jenni, L. & Winkler, R. (1994) *Moult and ageing of European Passerines*, London: Academic Press.

592 Kassara, C., Fric, J., Gschweng, M. & Sfenthourakis, S. 2012. Complementing the puzzle of  
593 Eleonora's Falcon (*Falco eleonora*) migration: new evidence from an eastern colony in the  
594 Aegean Sea. *Journal of Ornithology*, **153**: 839-848.

595 Lisovski, S. & Hahn, S. 2012. GeoLight-processing and analysing light-based geolocator data in R.  
596 *Methods in Ecology and Evolution*, **3**: 1055-1059.

597 Lisovski, S., Hewson, C. M., Klaassen, R. H. G., Korner-Nievergelt, F., Kristensen, M. W. & Hahn, S.  
 598 2012. Geolocation by light: accuracy and precision affected by environmental factors.  
 599 *Methods in Ecology and Evolution*, **3**: 603-612.

600 Martin, T. G., Chades, I., Arcese, P., Marra, P. P., Possingham, H. P. & Norris, D. R. 2007. Optimal  
 601 Conservation of Migratory Species. *Plos One*, **2**.

602 Müller, M., Spaar, R., Schifferli, L. & Jenni, L. 2005. Effects of changes in farming of subalpine  
 603 meadows on a grassland bird, the whinchat (*Saxicola rubetra*). *Journal of Ornithology*, **146**:  
 604 14-23.

605 Newton, I. (2008) *The migration ecology of birds*, Oxford: Academic Press.

606 R Development Core Team. 2014. R: A Language and Environment for Statistical Computing. Vienna,  
 607 Austria: R Foundation for Statistical Computing.

608 Reilly, J. R. & Reilly, R. J. 2009. Bet-hedging and the orientation of juvenile passerines in fall  
 609 migration. *Journal of Animal Ecology*, **78**: 990-1001.

610 Runge, C. A., Martini, T. G., Possingham, H. P., Willis, S. G. & Fuller, R. A. 2014. Conserving mobile  
 611 species. *Frontiers in Ecology and the Environment*, **12**: 395-402.

612 Runge, C. A., Watson, J. E. M., Butchart, S. H. M., Hanson, J. O., Possingham, H. P. & Fuller, R. A.  
 613 2015. Protected areas and global conservation of migratory birds. *Science*, **350**: 1255-1258.

614 Sillett, T. S. & Holmes, R. T. 2002. Variation in survivorship of a migratory songbird throughout its  
 615 annual cycle. *Journal of Animal Ecology*, **71**: 296-308.

616 Strandberg, R., Klaassen, R. H. G., Hake, M. & Alerstam, T. 2010. How hazardous is the Sahara  
 617 Desert crossing for migratory birds? Indications from satellite tracking of raptors. *Biology*  
 618 *Letters*, **6**: 297-300.

619 Studds, C. E., Kyser, T. K. & Marra, P. P. 2008. Natal dispersal driven by environmental conditions  
 620 interacting across the annual cycle of a migratory songbird. *Proceedings of the National*  
 621 *Academy of Sciences of the United States of America*, **105**: 2929-2933.

622 Sutherland, W. J. & Dolman, P. M. 1994. Combining behaviour and population dynamics with  
623 applications for predicting consequences of habitat loss. *Proceedings of the Royal Society of*  
624 *London Series B*, **255**: 133-138.

625 Taylor, C. M. & Norris, D. R. 2010. Population dynamics in migratory networks. *Theoretical Ecology*,  
626 **3**: 65-73.

627 Terraube, J., Mougeot, F., Cornulier, T., Verma, A., Gavrilov, A. & Arroyo, B. 2012. Broad wintering  
628 range and intercontinental migratory divide within a core population of the near-threatened  
629 pallid harrier. *Diversity and Distributions*, **18**: 401-409.

630 Tottrup, A. P., Klaassen, R. H. G., Strandberg, R., Thorup, K., Kristensen, M. W., Jorgensen, P. S.,  
631 Fox, J., Afanasyev, V., Rahbek, C. & Alerstam, T. 2012. The annual cycle of a trans-  
632 equatorial Eurasian-African passerine migrant: different spatio-temporal strategies for autumn  
633 and spring migration. *Proceedings of the Royal Society B-Biological Sciences*, **279**: 1008-  
634 1016.

635 Vickery, J. A., Ewing, S. R., Smith, K. W., Pain, D. J., Bairlein, F., Skorpilova, J. & Gregory, R. D.  
636 2014. The decline of Afro-Palaearctic migrants and an assessment of potential causes. *Ibis*,  
637 **156**: 1-22.

638 Webster, M. S., Marra, P. P., Haig, S. M., Bensch, S. & Holmes, R. T. 2002. Links between worlds:  
639 unraveling migratory connectivity. *Trends in Ecology & Evolution*, **17**: 76-83.

640

641

**Table 1.** Observed area and overlap (million square kilometres) of breeding locations compared to a random sample (lower and upper 95% CI), and mean distance ( $\pm$  1 SE km) between breeding locations for all birds (a) and first year birds (b) with 'best' summer sun elevation angle values (SEAs). See Supplementary Material Table S1 for a comparison of these values when using different sun elevation angle values: all are reasonably similar regardless of assumptions of sun elevation angle value.

	Overall	2013	2014	Difference 2014 - 2013	Lower 95% CI value for the random sample	Upper 95% CI value for the random sample	P value
<b>a. All birds</b>	<i>N</i> = 29	<i>N</i> = 12	<i>N</i> = 17				
Area	2.549	0.651	2.157	1.506	-0.612	1.225	0.01
Overlap	14.7				32.4	76.2	0.01
Mean distance		604 $\pm$ 18	869 $\pm$ 33	$t_{28} = 5.2$			0.0001
<b>b. First year birds</b>	<i>N</i> = 13	<i>N</i> = 7	<i>N</i> = 6				
Area	1.700	0.470	1.199	0.729	-0.857	0.589	0.02
Overlap	7.9				3.8	56.8	0.06
Mean distance		598 $\pm$ 40	1059 $\pm$ 72	$t_{12} = 4.1$			0.0010



**Table 2.** Observed area and overlap (million square kilometres) of breeding locations compared to a random sample (lower and upper 95% CI), and mean distance ( $\pm$  1 SE km) between breeding locations for all birds (a.) and first year birds (b.) with ‘best’ summer sun elevation angle values (SEAs), latitudes adjusted for the effect of no light stalks on four birds in 2014. Latitude adjustments for each analysis are given in italics. Sample sizes as in Table 1. See Supplementary Material Table S2 for a comparison when using different sun elevation angle values: all are reasonably similar regardless of assumptions of sun elevation angle value.

	Overall	2013	2014	Difference 2014 - 2013	Lower 95% CI value for the random sample	Upper 95% CI value for the random sample	<i>P</i> value
<b>a. All birds</b>	<i>Light stalk adjustment for latitude <math>-4.45 \pm 1.6</math>, <math>F_{1,26} = -2.7</math>, <math>P = 0.010</math></i>						
Area	1.681	0.651	1.650	0.999	-0.440	0.814	0.01
Overlap	36.5				35.2	77.8	0.06
Mean distance		$604 \pm 18$	$820 \pm 29$	$t_{28} = 4.5$			0.0001
<b>b. First year birds</b>	<i>Adjustment as for all birds above</i>						
Area	1.275	0.470	1.232	0.762	-0.651	0.486	0.01
Overlap	33.2				4.2	60.3	0.51
Mean distance		$598 \pm 40$	$1042 \pm 63$	$t_{12} = 4.3$			0.001

## Figure legends

**Figure 1.** How sampling at any one point on the non-breeding ground can result in sampling birds over a large breeding range (i.e. low connectivity arises from wide migratory dispersal). This model assumes high natal (see Greenwood & Harvey 1982) and non-breeding (see Cresswell 2014) site fidelity. The model also assumes that the non-breeding range of a migrant is at a very large scale (as it is with the Whinchat and most other Afro-Palearctic migrants which have non-breeding ranges that encompass much of Africa), where migrants may have small, concentrated, non-breeding areas thus tagging from these areas will also result in sampling birds over a large breeding range.

**Figure 2.** How sampling at any one point on the wintering ground in different years can result in sampling birds with different breeding ranges (i.e. inter-annual variation in connectivity arises through stochastic cohort effects such as, for example wind conditions during migration varying between years). This model assumes high natal site fidelity (see Greenwood & Harvey 1982) and high non-breeding site fidelity if a first year bird has found a non-breeding site that allowed survival (see the serial residency hypothesis Cresswell 2014).

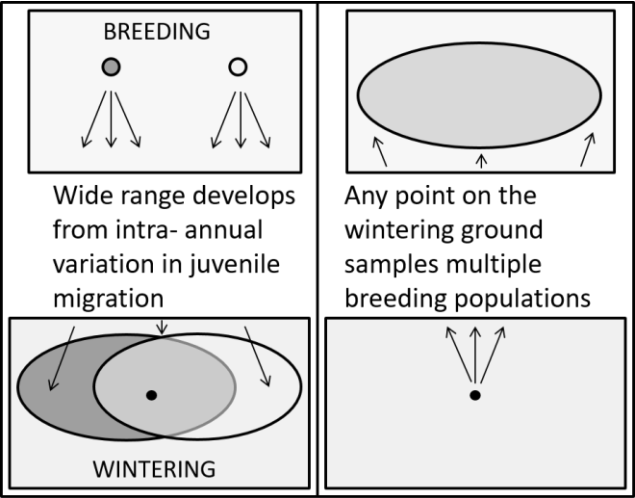
**Figure 3.** Breeding locations and range for 29 Whinchats tagged at Jos Nigeria (A) in 2013 (black circular points, white fill) and in 2014 (white triangular points and darker grey fill). Each point is plotted at the mean of all latitudes and longitudes recorded over the least variable 2 week stationary period during mid-May – mid-July;  $\pm 2$  Standard Errors from the mean location are also plotted but only the highly variable locations are clearly visible at this scale. All birds plotted at the ‘best’ estimate locations, unadjusted for light stalk absence (see Table 1).

**Figure 4.** How range depended on sample size. Mean range plotted for 1000 random subsamples of points, each of different sample size. a. All birds and years pooled, unadjusted for light stalk absence. A linear fit is plotted with 95% CI plotted as dotted lines (range =  $(0.097 \pm 0.003 * (\text{sample size})) + (-0.03 \pm 0.04)$ ,  $F_{1,12} = 37.2$ ,  $P < 0.0001$ ;  $R^2 = 0.99$ ;  $\Delta\text{AIC} = + 0.2$ , compared to a quadratic model including the square of sample size which gives a near identical plot). b. All birds and years pooled, adjusted for light stalk absence. Line of best fit is quadratic, 95% CI plotted as dotted lines (range =  $(0.094 \pm 0.005 * (\text{sample size})) + (-0.0011 \pm 0.0002 * (\text{sample size}^2)) + (-0.04 \pm 0.03)$ ,  $F_{1,11} = 1010$ ,  $P < 0.0001$ ;  $R^2 = 0.99$ ;  $\Delta\text{AIC} = - 19.8$  better than a linear model without the square of sample size). c. All birds, years plotted separately, unadjusted for light stalk absence. Lines of best fit are linear (1 standard error plotted as dotted lines) and the gradients of the two lines are highly significantly

different. d. All birds, years plotted separately, adjusted for light stalk absence. Lines of best fit are linear (1 standard error plotted as dotted lines) and the gradients of the two lines are highly significantly different.

**Figure 5.** Breeding locations and range for Whinchats tagged in 2013 (black circular points, white fill, N = 12) and in 2014 (white triangular points and darker grey fill, N = 17). Each point is plotted at the mean of all latitudes and longitudes recorded over the least variable 2 week stationary period during mid-May – mid-July (N = 28 locations), +/- 2 Standard Errors from the mean location are also plotted but only the highly variable locations are visible at this scale. a. All birds plotted at the 'best' estimate locations, unadjusted for light stalk absence (i.e. an enlarged version of Fig. 3). b. All birds plotted at the 'best' estimate locations, adjusted for light stalk absence. c. First year birds plotted at the 'best' estimate locations, unadjusted for light stalk absence. d. First year birds plotted at the 'best' estimate locations, adjusted for light stalk absence. Ranges that arose from varying the assumptions with respect to sun elevation angles (see Table 1 and 2) are broadly similar and are illustrated in Supplementary Material Figures S3 – S20).

704 Figure 1:

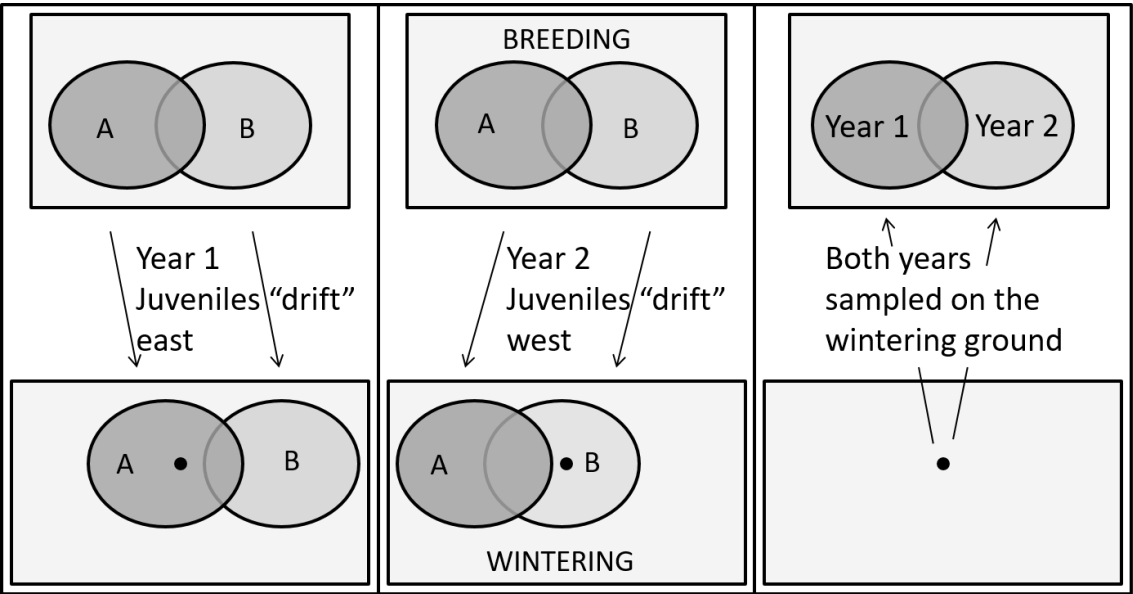


705

706

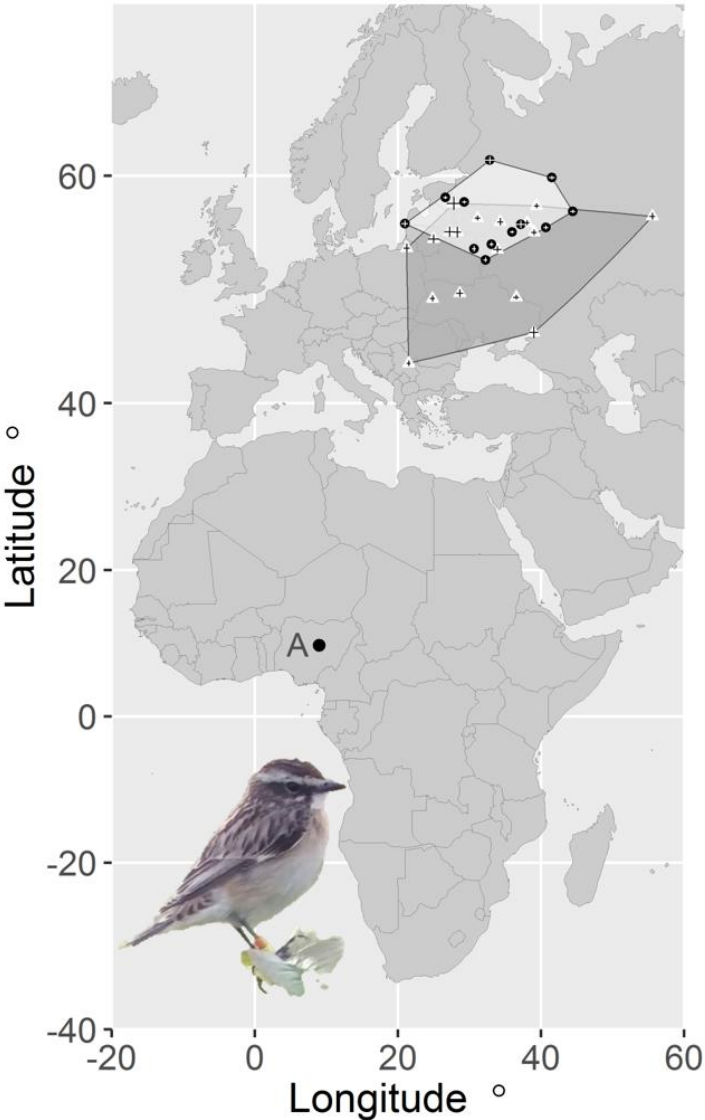
707

708 Figure 2:



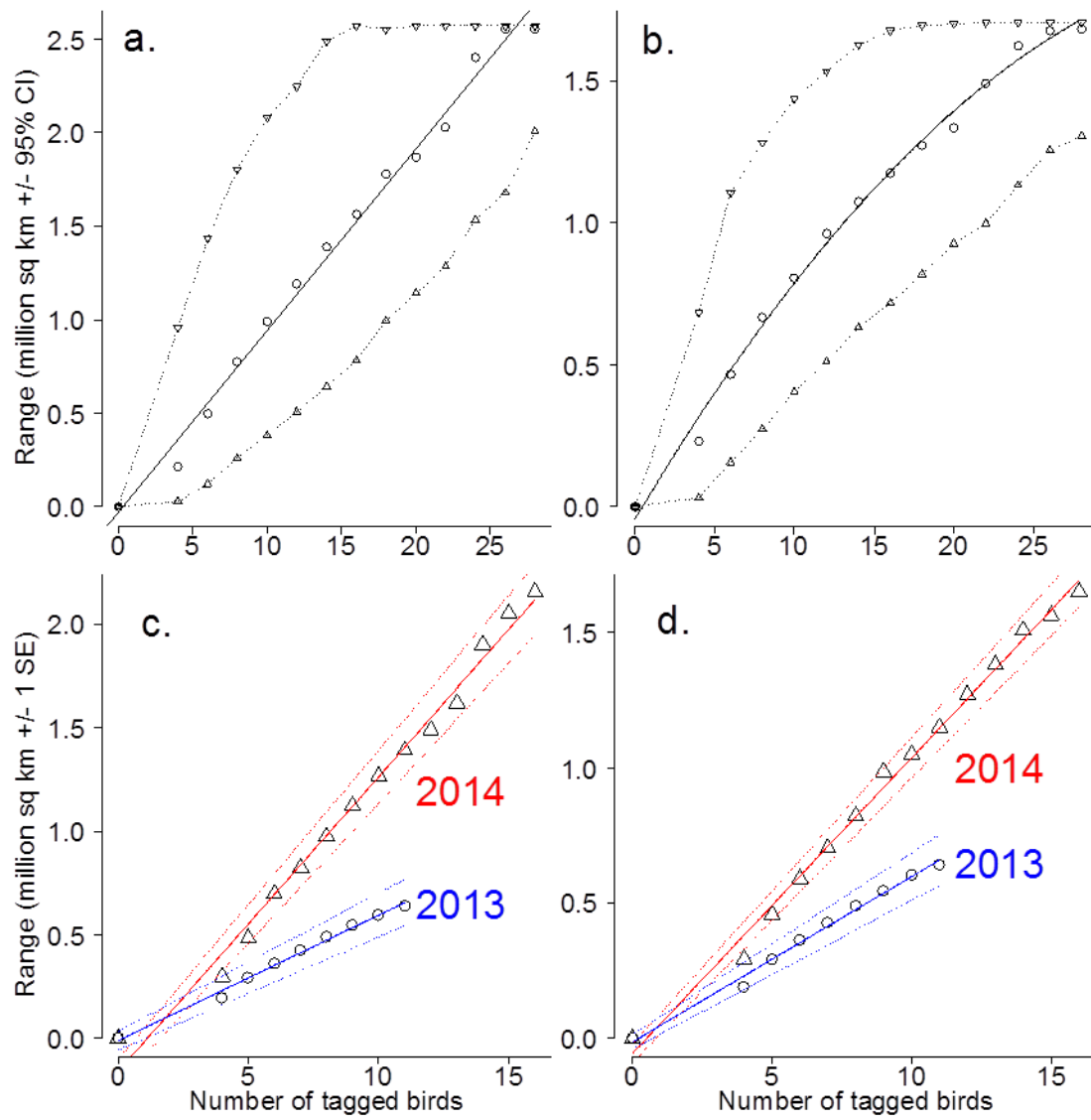
709

710



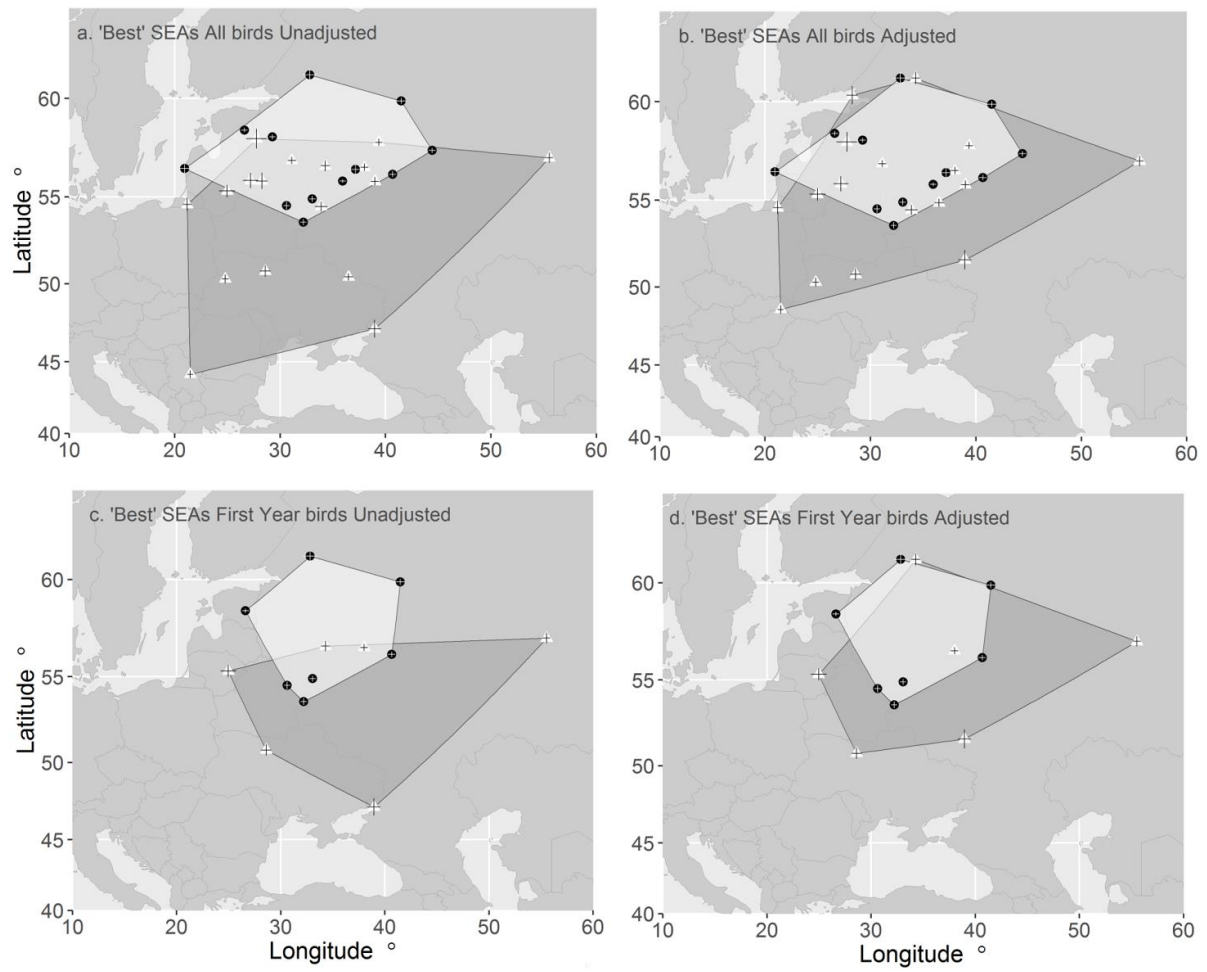
712  
713

714 Figure 4



715

Figure 5





721 **ONLINE ONLY SUPPLEMENTARY MATERIAL**

722

723 **Low and annually variable migratory connectivity in a long distance migrant:**  
724 **Whinchats *Saxicola rubetra* may show a bet-hedging strategy**

725

726 EMMA BLACKBURN<sup>1,2</sup>, MALCOLM BURGESS<sup>3</sup>, BENEDICTUS FREEMAN<sup>2,6</sup>, ALICE RISELY<sup>4</sup>, ARIN  
727 IZANG<sup>2</sup>, SAM IVANDE<sup>1,2</sup>, CHRIS HEWSON<sup>5</sup> & WILL CRESSWELL<sup>1\*,2</sup>

728

729 <sup>1</sup>*Centre for Biological Diversity, University of St Andrews, Harold Mitchell Building, St Andrews, Fife*  
730 *KY16 9TH, UK*

731 <sup>2</sup>*AP Leventis Ornithological Research Institute, Jos, Nigeria*

732 <sup>3</sup>*RSPB Centre for Conservation Science, The Lodge, Sandy, Beds SG19 2DL, UK*

733 <sup>4</sup>*Centre for Integrative Ecology, Deakin University, Geelong, VIC 3216, Australia*

734 <sup>5</sup>*British Trust for Ornithology, The Nunnery, Thetford, IP24 2PU UK*

735 <sup>6</sup>*Fauna & Flora International (FFI), 2 Harmon's Compound, Oldest Congo Town, Monrovia, Liberia*

736 \*Correspondence author: wrlc@st-and.ac.uk

737

**Table S1.** Observed area and overlap (million square kilometres) of breeding locations compared to a random sample (lower and upper 95% CI: random low and random up respectively in table), and mean distance ( $\pm 1$  SE km) between breeding locations for all birds (a) and first year birds (b) with different values of sun elevation angle (SEA).

<b>a. ALL</b>	Overall	2013	2014	Diff	Random low	Random up	P value
<b>1. BEST SUMMER SEA</b>	<i>N</i> = 29	<i>N</i> = 12	<i>N</i> = 17				
Area	2.549	0.651	2.157	1.506	-0.612	1.225	0.01
Overlap	14.7				32.4	76.2	0.01
Mean distance		604 $\pm$ 18	869 $\pm$ 33	$t_{28} = 5.2$			0.0001
<b>2. WINTER SEA</b>							
Area	2.332	0.703	2.230	1.527	-0.586	1.063	0.01
Overlap	27.1				34.3	77.9	0.01
Mean distance		631 $\pm$ 20	859 $\pm$ 34	$t_{28} = 4.2$			0.0002
<b>3. AVERAGE SEA</b>							
Area	2.514	0.744	2.183	1.439	-0.566	1.215	0.02
Overlap	20.9				35.3	77.1	0.01
Mean distance		634 $\pm$ 19	862 $\pm$ 33	$t_{28} = 4.4$			0.0002
<b>4. SUMMER SEA JUNE</b>							
Area	2.519	0.603	2.049	1.446	-0.567	1.147	0.01
Overlap	20.9				35.3	75.2	0.01
Mean distance		593 $\pm$ 18	903 $\pm$ 32	$t_{28} = 6.4$			0.0001
<b>5. MINIMUM RANGE SUMMER SEA</b>							
Area	2.351	0.503	1.971	1.468	-0.568	1.144	0.01
Overlap	8.7				32.1	75.9	0.01
Mean distance		591 $\pm$ 19	839 $\pm$ 33	$t_{28} = 4.8$			0.0001
<b>6. MAXIMUM OVERLAP SUMMER SEA</b>							
Area	2.351	0.640	2.126	1.486	-0.520	1.061	0.01
Overlap	21.9				34.0	76.6	0.01
Mean distance		599 $\pm$ 18	870 $\pm$ 33	$t_{28} = 5.3$			0.0001
<b>b. FIRST YEARS</b>							
<b>1. BEST SUMMER SEA</b>	<i>N</i> = 13	<i>N</i> = 7	<i>N</i> = 6				
Area	1.700	0.470	1.199	0.729	-0.857	0.589	0.02
Overlap	7.9				3.8	56.8	0.06
Mean distance		598 $\pm$ 40	1059 $\pm$ 72	$t_{12} = 4.1$			0.0010
<b>2. WINTER SEA</b>							
Area	1.462	0.487	1.145	0.658	-0.706	0.590	0.04
Overlap	18.9				4.8	58.6	0.24
Mean distance		620 $\pm$ 43	1032 $\pm$ 73	$t_{12} = 3.6$			0.003
<b>3. AVERAGE SEA</b>							
Area	1.689	0.518	1.199	0.681	-0.832	0.622	0.05
Overlap	11.6				3.9	60.2	0.10
Mean distance		640 $\pm$ 41	1060 $\pm$ 72	$t_{12} = 3.7$			0.003
<b>4. SUMMER SEA JUNE</b>							
Area	1.672	0.429	1.153	0.681	-0.78	0.577	0.03
Overlap	2.7				4.6	59.3	0.04
Mean distance		567 $\pm$ 40	1013 $\pm$ 70	$t_{12} = 3.3$			0.007

**Table S2.** Observed area and overlap (million square kilometres) of breeding locations compared to a random sample (lower and upper 95% CI: random low and random up respectively in table), and mean distance ( $\pm 1$  SE km) between breeding locations for all birds (a.) and first year birds (b.) with different values of SEA, latitudes adjusted for the effect of no light stalks on four birds in 2014. Latitude adjustments for each SEA value analysis are given in italics. Sample sizes as in Table 1.

<b>a. ALL</b>	Overall	2013	2014	Diff	Random low	Random up	P value
<b>1. BEST SUMMER SEA</b>	<i>Light stalk adjustment for latitude <math>-4.45 \pm 1.6</math>, <math>F_{1,26} = -2.7</math>, <math>P = 0.010</math></i>						
Area	1.681	0.651	1.650	0.999	-0.440	0.814	0.01
Overlap	36.5				35.2	77.8	0.06
Mean distance		$604 \pm 18$	$820 \pm 29$	$t_{28} = 4.5$			0.0001
<b>2. WINTER SEA</b>	<i>Light stalk adjustment for latitude <math>-3.75 \pm 1.7</math>, <math>F_{1,26} = -2.2</math>, <math>P = 0.034</math></i>						
Area	1.737	0.703	1.727	1.024	-0.523	0.871	0.02
Overlap	41.1				35.2	77.7	0.07
Mean distance		$631 \pm 20$	$828 \pm 30$	$t_{28} = 3.9$			0.0005
<b>3. AVERAGE SEA</b>	<i>Light stalk adjustment for latitude <math>-4.31 \pm 1.7</math>, <math>F_{1,26} = -2.6</math>, <math>P = 0.014</math></i>						
Area	1.754	0.744	1.721	0.977	-0.385	0.869	0.02
Overlap	41.7				34.4	77.7	0.07
Mean distance		$634 \pm 19$	$817 \pm 29$	$t_{28} = 3.8$			0.0007
<b>4. SUMMER SEA JUNE</b>	<i>Light stalk adjustment for latitude <math>-5.02 \pm 1.5</math>, <math>F_{1,26} = -3.4</math>, <math>P = 0.002</math></i>						
Area	1.835	0.603	1.814	1.211	-0.454	0.897	0.01
Overlap	31.1				35.2	76.0	0.03
Mean distance		$593 \pm 18$	$824 \pm 30$	$t_{28} = 4.7$			0.0001
<b>b. FIRST YEARS</b>							
<b>1. BEST SUMMER SEA</b>	<i>Adjustment as for 1. above</i>						
Area	1.275	0.470	1.232	0.762	-0.651	0.486	0.01
Overlap	33.2				4.2	60.3	0.51
Mean distance		$598 \pm 40$	$1042 \pm 63$	$t_{12} = 4.3$			0.001
<b>2. WINTER SEA</b>	<i>Adjustment as for 2. above</i>						
Area	1.22	0.487	1.176	0.689	-0.605	0.442	0.01
Overlap	39.4				3.6	58.8	0.71
Mean distance		$620 \pm 43$	$1028 \pm 66$	$t_{12} = 3.8$			0.0020
<b>3. AVERAGE SEA</b>	<i>Adjustment as for 3. above</i>						
Area	1.327	0.518	1.227	0.709	-0.66	0.449	0.01
Overlap	33.4				2.2	60.1	0.52
Mean distance		$640 \pm 41$	$1040 \pm 63$	$t_{12} = 3.8$			0.002
<b>4. SUMMER SEA JUNE</b>	<i>Adjustment as for 4. above</i>						
Area	1.348	0.429	0.829	0.400	-0.652	0.521	0.12
Overlap	3.3				4.1	58.5	0.04
Mean distance		$567 \pm 40$	$950 \pm 65$	$t_{12} = 3.1$			0.01

**Table S3:** Full models (left) to investigate whether there was any bias in latitude due to age, sex, absence of light stalks, SEA value used (if varied in a model, i.e. not when average SEA value was used for all birds), and whether the SEA value used was estimated using the H-E calibration or an average (if this was relevant in a model), controlling for year. Final models (right) with the only significant term retained (light stalk absence) controlling for year to determine the final adjustment in latitude due to four tags in 2014 having no light stalks.

	Full Model				Final Model			
<b>BEST SUMMER SEA</b>	Est.	Std. Error	t value	P value	Est.	Std. Error	t value	P value
(Intercept) Adult Female 2013	55.7	6.2	9.0	<0.0001	56.9	0.88	64.9	<0.0001
Age first year	-0.88	1.9	-0.5	0.64				
Sex Male	-0.25	2.2	-0.1	0.91				
Year 2014	2.4	12.9	0.2	0.85	-1.6	1.2	-1.3	0.20
With light stalk	-4.3	2.2	-2.0	0.067	-4.5	1.6	-2.7	0.10
Summer SEA	-0.48	1.7	-0.3	0.77				
Average SEA used YES	0.79	2.3	0.3	0.74				
Year 2014 * Summer SEA	1.4	3.2	0.4	0.67				
Year 2014 * Av SEA YES	1.3	3.1	0.4	0.67				
	$F_{8,17} = 1.4, P = 0.27, R^2 = 0.11$				$F_{2,26} = 7.1, P = 0.004, R^2 = 0.30$			
<b>WINTER SEA</b>	Est.	Std. Error	t value	P value	Est.	Std. Error	t value	P value
(Intercept) Adult Female 2013	41.4	14.4	2.9	0.009	55.3	0.9	60.7	<0.0001
Age first year	0.49	1.6	0.3	0.76				
Sex Male	0.15	2.1	0.1	0.94				
Year 2014	26.5	17.6	1.5	0.15	-0.31	1.3	-0.2	0.81
With light stalk	-5.0	2.3	-2.2	0.040	-3.7	1.7	-2.2	0.034
Winter SEA	-2.7	2.8	-1.0	0.35				
Year 2014 * Winter SEA	5.9	3.8	1.5	0.14				
	$F_{6,19} = 1.3, P = 0.30, R^2 = 0.07$				$F_{2,26} = 3.2, P = 0.057, R^2 = 0.14$			
<b>AVERAGE SEA</b>	Est.	Std. Error	t value	P value	Est.	Std. Error	t value	P value
(Intercept) Adult Female 2013	56.4	2.7	20.6	<0.0001	56.3	0.9	62.7	<0.0001
Age first year	0.16	1.6	0.1	0.92				
Sex Male	-0.20	2.1	-0.1	0.92				
Year 2014	-1.0	1.6	-0.6	0.55	-0.97	1.3	-0.8	0.45
With light stalk	-4.4	2.1	-2.1	0.044	-4.3	1.7	-2.6	0.015
	$F_{4,21} = 2.1, P = 0.11, R^2 = 0.15$				$F_{2,26} = 5.2, P = 0.012, R^2 = 0.23$			
<b>SUMMER SEA JUNE</b>	Est.	Std. Error	t value	P value	Est.	Std. Error	t value	P value
(Intercept) Adult Female 2013	52.8	5.4	9.7	<0.0001	56.7	0.8	71.5	<0.0001
Age first year	-0.52	1.6	-0.32	0.75				
Sex Male	2.0	2.0	1.0	0.32				

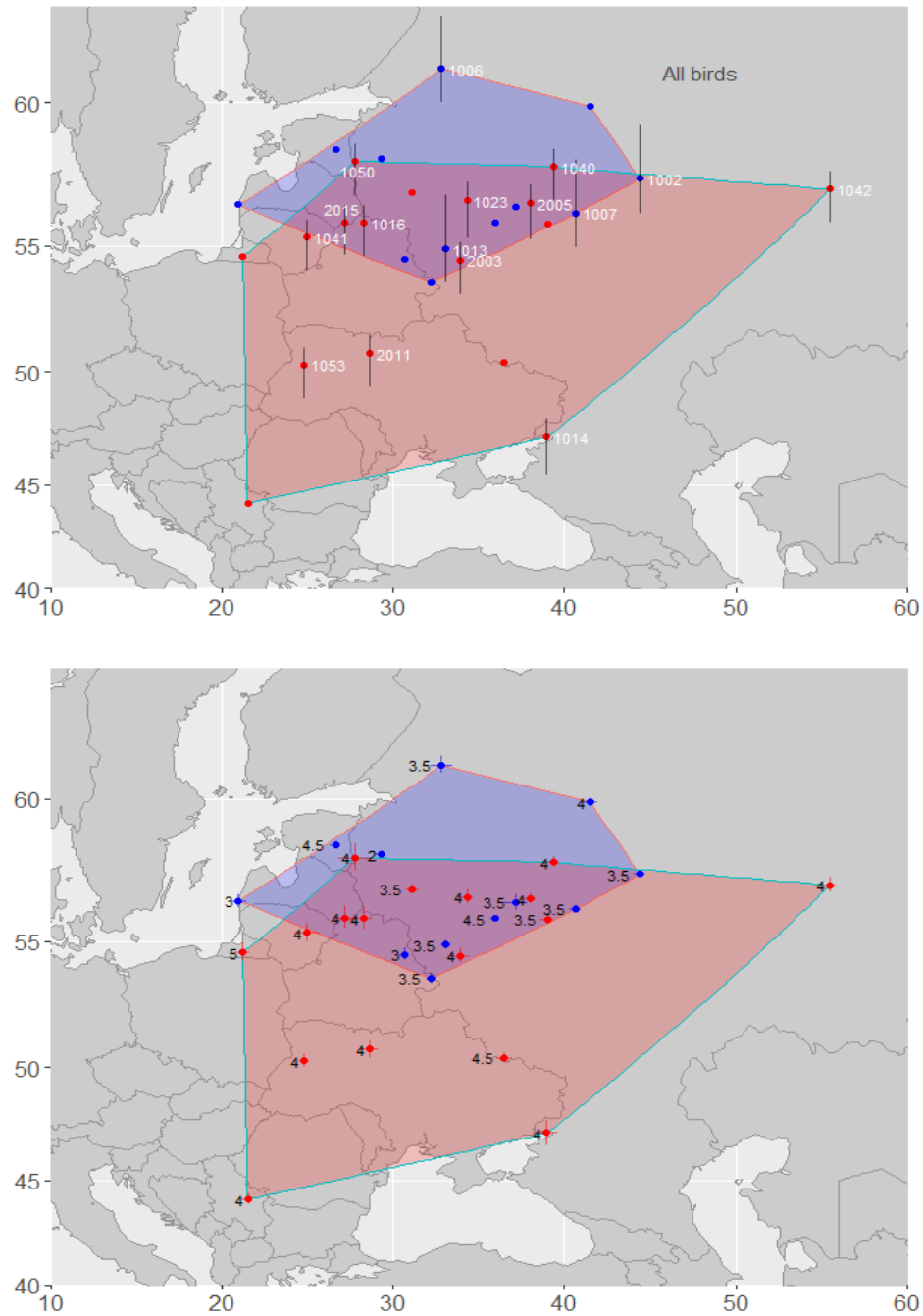
Year 2014	2.3	11.4	0.2	0.84	-2.6	1.1	-2.3	0.030
With light stalk	-3.9	1.9	-2.1	0.055	-5.0	1.5	-3.4	0.002
Summer SEA	-0.53	1.5	-0.4	0.72				
Average SEA used YES	1.3	2.0	0.6	0.52				
Year 2014 * Summer SEA	1.3	2.9	0.5	0.65				
Year 2014 * Av SEA YES	-0.44	2.7	-0.2	0.88				
$F_{8,17} = 2.8, P = 0.038, R^2 = 0.36$					$F_{2,26} = 13.6, P < 0.0001, R^2 = 0.47$			

755

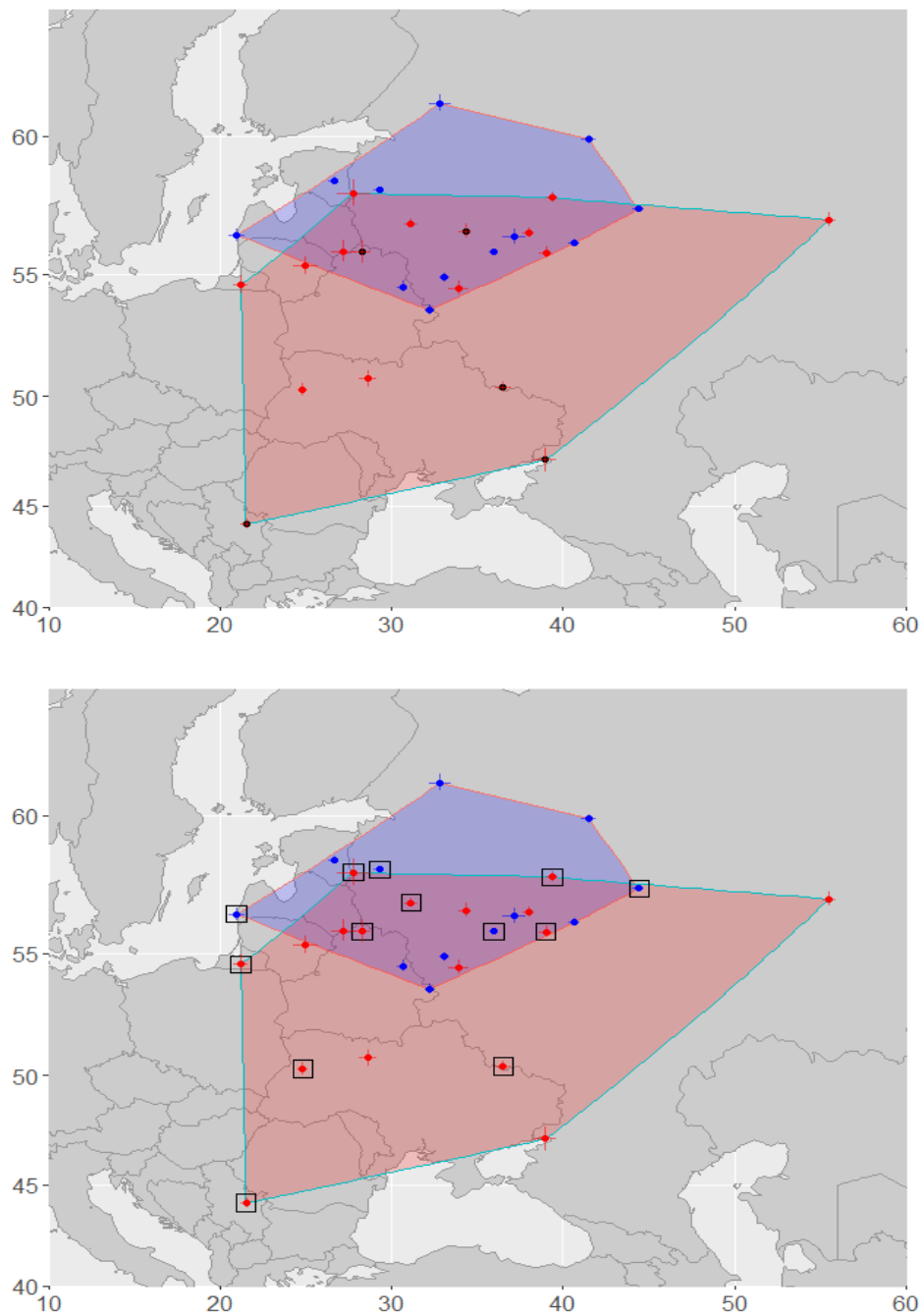
756 **Table S4:** Models to investigate whether there was any bias in longitude due to age, sex, absence of  
757 light stalks, SEA value used (if varied in a model, i.e. not when average SEA value was used for all  
758 birds), and whether the SEA value used was estimated using the H-E calibration or an average (if this  
759 was relevant in a model), controlling for year.

<b>BEST SUMMER SEA</b>	Est.	Std. Error	t value	P value
(Intercept) Adult Female 2013	21.7	14.9	1.5	0.17
Age first year	3.0	4.5	0.7	0.52
Sex Male	0.19	5.4	0.04	0.97
Year 2014	33.2	31.3	1.1	0.30
With light stalk	-0.27	5.3	-0.1	0.96
Summer SEA	-2.1	4.0	-0.5	0.60
Average SEA used YES	6.2	5.6	1.1	0.28
Year 2014 * Summer SEA	8.2	7.9	1.0	0.31
Year 2014 * Av SEA YES	-4.4	7.5	-0.6	0.57
$F_{8,17} = 0.5, P = 0.82, R^2 = 0$				
<b>WINTER SEA</b>	Est.	Std. Error	t value	P value
(Intercept) Adult Female 2013	48.6	35.9	1.4	0.19
Age first year	4.1	4.1	1.0	0.33
Sex Male	0.25	5.2	0.05	0.96
Year 2014	1.8	44.0	0.04	0.97
With light stalk	-2.9	5.7	-0.5	0.61
Winter SEA	3.7	7.1	0.5	0.61
Year 2014 * Winter SEA	0.82	9.5	0.1	0.93
$F_{6,19} = 0.4, P = 0.85, R^2 = 0$				
<b>AVERAGE SEA</b>	Est.	Std. Error	t value	P value
(Intercept) Adult Female 2013	30.6	6.7	4.6	0.0002
Age first year	4.5	3.9	1.2	0.26
Sex Male	0.033	5.0	0.01	0.99
Year 2014	0.64	3.9	0.2	0.87
With light stalk	-1.1	5.0	-0.2	0.83
$F_{4,21} = 0.5, P = 0.75, R^2 = 0$				
<b>SUMMER SEA JUNE</b>	Est.	Std. Error	t value	P value
(Intercept) Adult Female 2013	20.9	15.4	1.4	0.19
Age first year	3.7	4.7	0.8	0.44
Sex Male	1.2	5.5	0.2	0.84
Year 2014	33.5	32.3	1.0	0.31
With light stalk	-0.25	5.4	-0.05	0.96
Summer SEA	-2.0	4.1	-0.5	0.64
Average SEA used YES	6.2	5.8	1.1	0.30
Year 2014 * Summer SEA	8.3	8.2	1.0	0.32
Year 2014 * Av SEA YES	-5.1	7.8	-0.7	0.52
$F_{8,17} = 0.5, P = 0.81, R^2 = 0$				

**Figure S1:** Top - Loggers where an average summer SEA was used labelled in white, showing the range of SEAs (black vertical lines) for all other loggers where the calibration was successful for that year. Bottom – the SEA (sun elevation angle used for each logger in the “best” Analysis 1. Blue 2013, Red 2014.



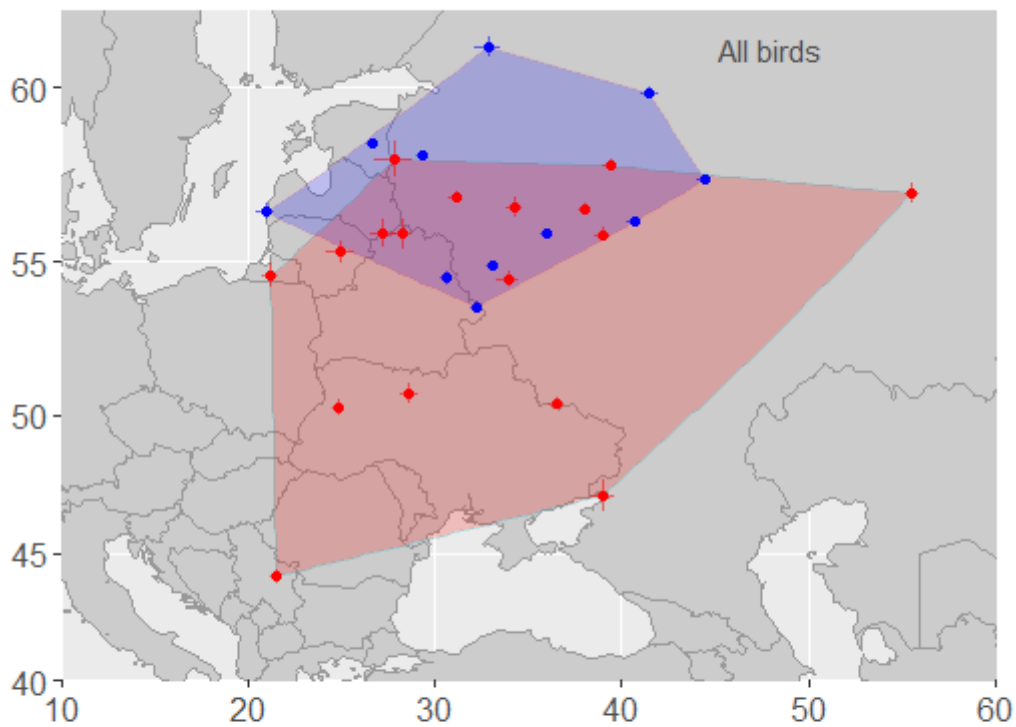
**Figure S2:** Top - Loggers in 2014 without a light stalk plotted in black (and so where breeding location latitudes were adjusted northwards according to the models in Table S1). Bottom - Loggers of adult birds plotted within black squares





**Figure S3:** All birds summer unadjusted (*BEST SUMMER SEA* Table S1.a.1). Blue 2013, Red 2014; +/- 2 standard errors in latitude and longitude are plotted for each location.

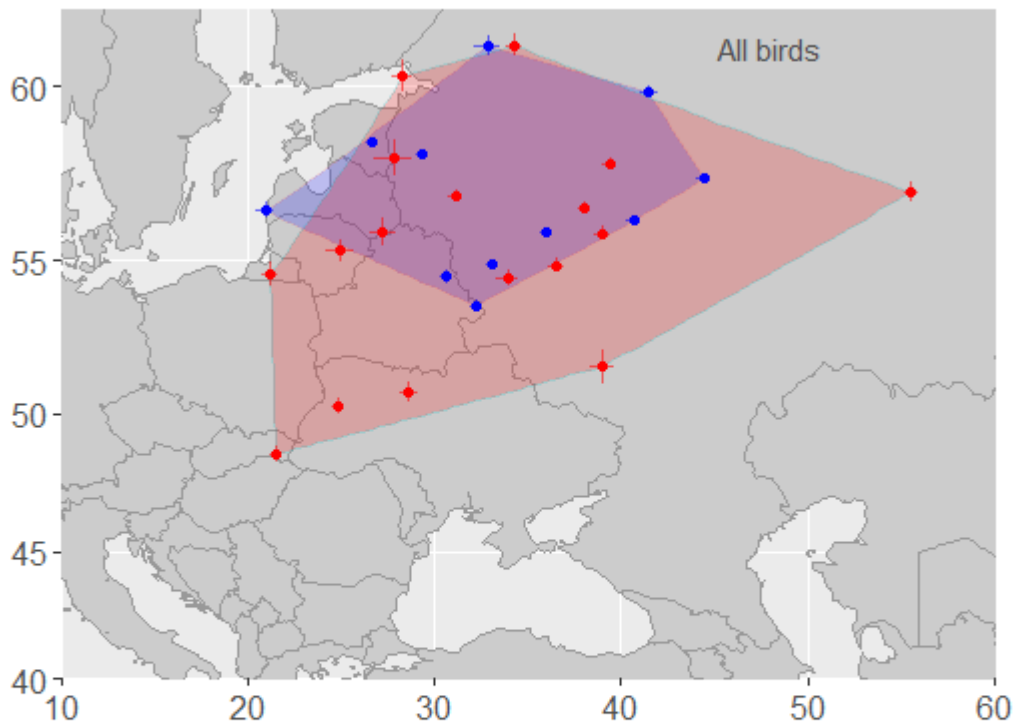
2013 0.651; 2014 2.157 million square km (mskm)  
 604 (+/- 18) vs 869 (+/- 33) km ( $t_{28} = 5.2$ ,  $P = 0.0001$ )  
 Area diff = 1.506 mskm (99th percentile)  
 Random -0.612 to 1.225 mskm  
 Overall 2.549 mskm = 26.2% Europe land area  
 14.7% (99th percentile)  
 Random 32.4% - 76.2%



**Figure S4:** All birds summer adjusted (*BEST SUMMER SEA* Table S2.a.1). Blue 2013, Red 2014; +/- 2

standard errors in latitude and longitude are plotted for each location.

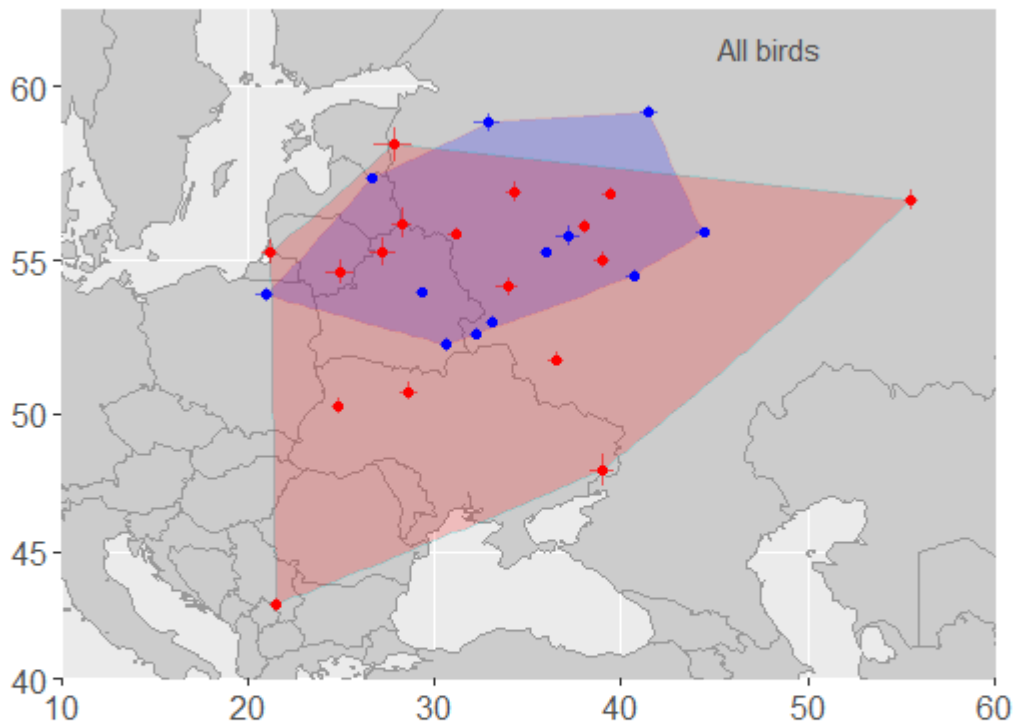
2013 0.651; 2014 1.650 million square km  
604 (+/- 18) vs 820 (+/- 29) km ( $t_{28} = 4.5$ ,  $P = 0.0001$ )  
Area diff = 0.999 mskm (99<sup>th</sup> percentile)  
Random -0.440 to 0.814 mskm  
Overall 1.681 mskm = 16.5%  
37.9% (94<sup>th</sup> percentile)  
Random 35.2% - 77.8%



**Figure S5:** All birds winter unadjusted (*WINTER SEA* Table S1.a.2). Blue 2013, Red 2014; +/- 2

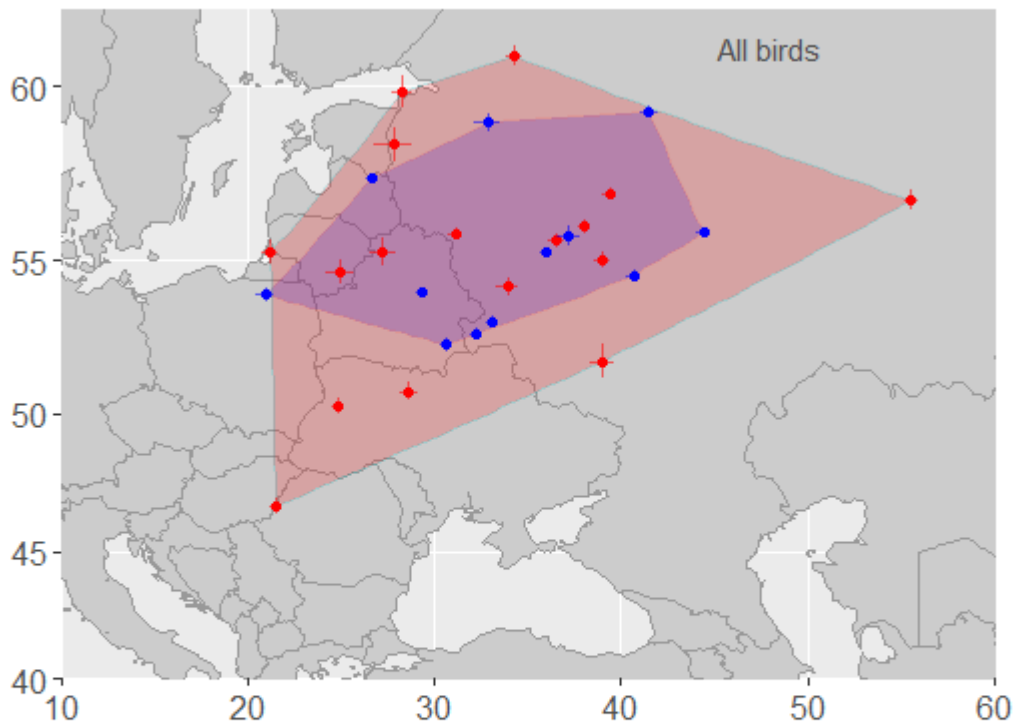
standard errors in latitude and longitude are plotted for each location.

2013 0.703; 2014 2.230 million square km  
 631 (+/- 20) vs 859 (+/- 34) km ( $t_{28} = 4.2$ ,  $P = 0.0002$ )  
 Area diff = 1.527 mskm (99<sup>th</sup> percentile)  
 Random -0.586 to 1.063 mskm  
 Overall 2.332 mskm = 22.9%  
 27.1% (99<sup>th</sup> percentile)  
 Random 34.3% - 77.9%



**Figure S6:** All birds winter adjusted (*WINTER SEA* Table S2.a.2). Blue 2013, Red 2014; +/- 2 standard errors in latitude and longitude are plotted for each location.

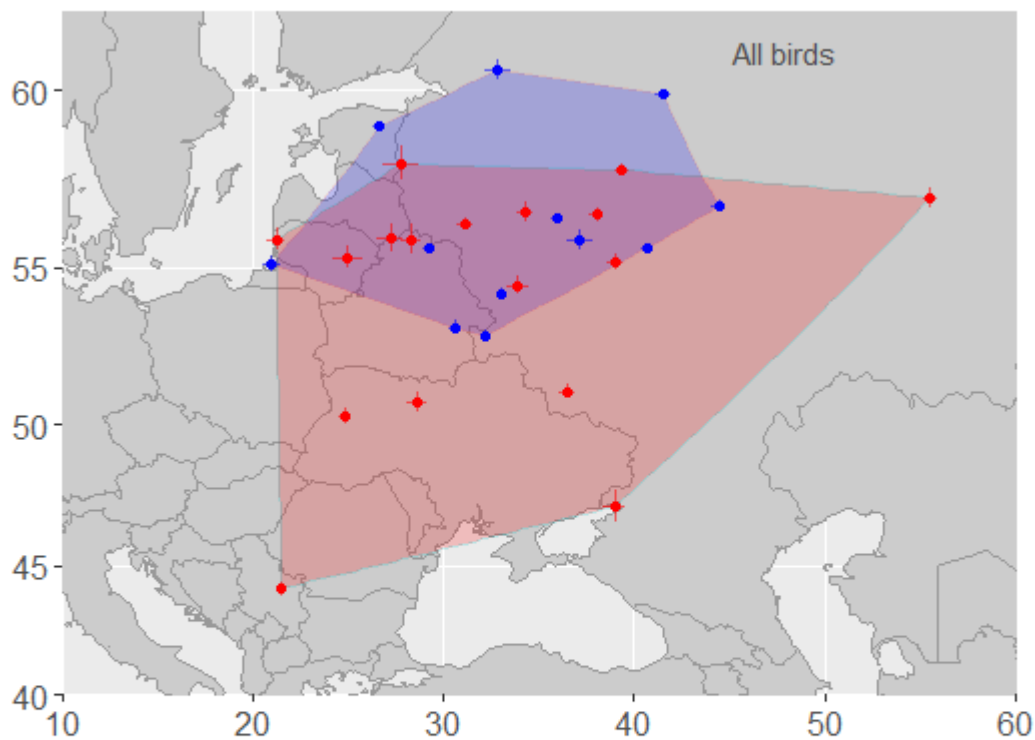
2013 0.703; 2014 1.727 million square km  
 631 (+/- 20) vs 828 (+/- 30) km ( $t_{28} = 3.9$ ,  $P = 0.0005$ )  
 Area diff = 1.024 mskm (98<sup>th</sup> percentile)  
 Random -0.523 to 0.871  
 Overall 1.737 mskm = 17.0%  
 41.1% (93rd percentile)  
 Random 35.2% - 77.7%



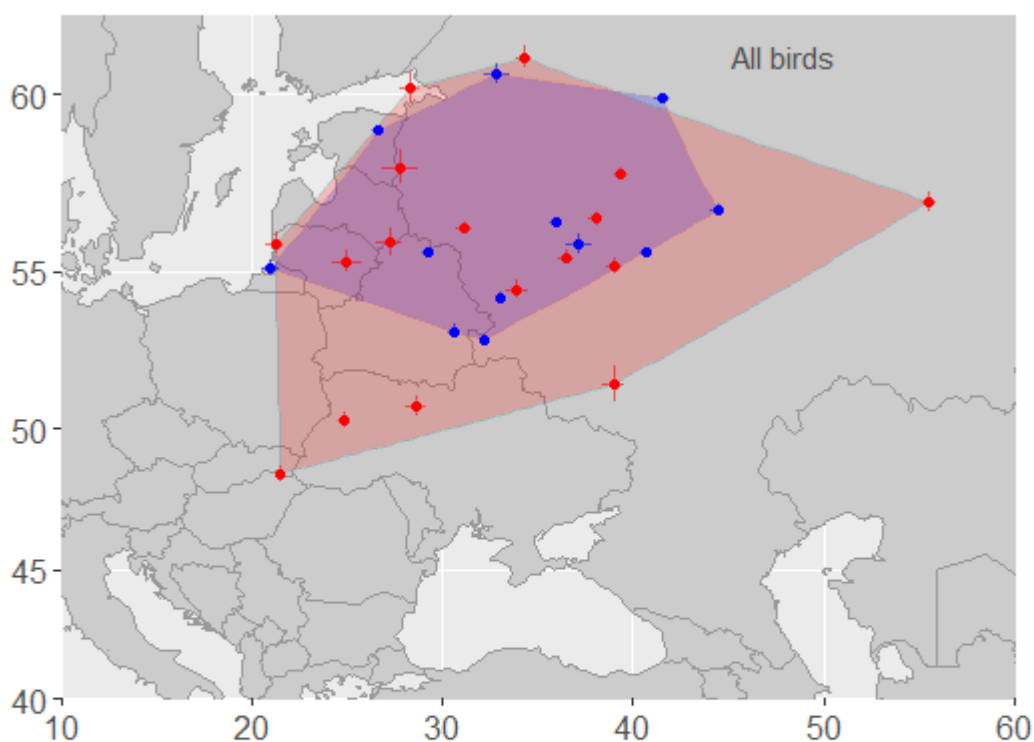
**Figure S7:** All birds average of SEA angles from 2 to 6, unadjusted (*AVERAGE SEA* Table S1.a.3). Blue

2013, Red 2014;  $\pm 2$  standard errors in latitude and longitude are plotted for each location.

2013 0.744; 2014 2.183 million square km  
 634 ( $\pm 19$ ) vs 862 ( $\pm 33$ ) km ( $t_{28} = 4.4$ ,  $P = 0.0002$ )  
 Area diff = 1.439 mskm (98<sup>th</sup> percentile)  
 Random -0.566 to 1.215 mskm  
 Overall 2.514 mskm = 24.7%  
 20.9% (<99<sup>th</sup> percentile)  
 Random 35.3% - 77.1%



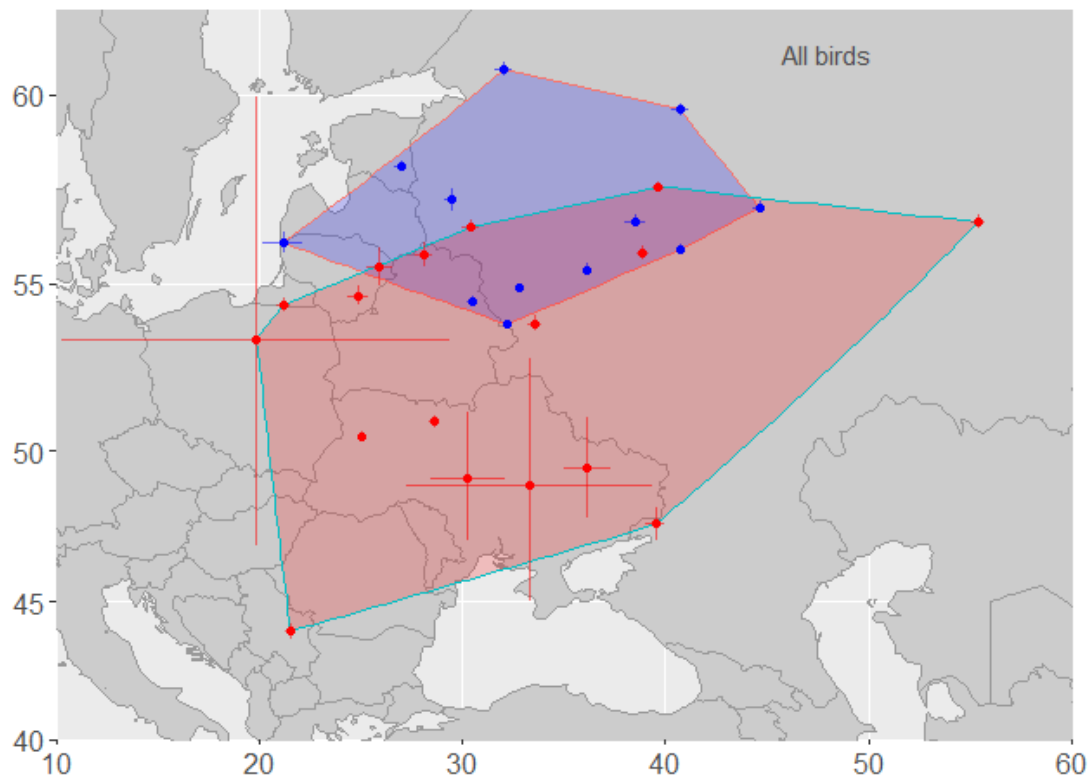
823 **Figure S8:** All birds average from 2 to 6, adjusted (*AVERAGE SEA* Table S1.a.3). Blue 2013, Red 2014;  
 824 +/- 2 standard errors in latitude and longitude are plotted for each location.  
 825 2013 0.744; 2014 1.721 million square km  
 826 634 (+/- 19) vs 817 (+/- 29) km ( $t_{28} = 3.8$ ,  $P = 0.0007$ )  
 827 Area diff = 0.977 mskm (98<sup>th</sup> percentile)  
 828 Random -0.385 to 0.869 mskm  
 829 Overall 1.754 mskm = 17.2%  
 830 41.7% (93<sup>rd</sup> percentile)  
 831 Random 34.4% - 77.7%



832

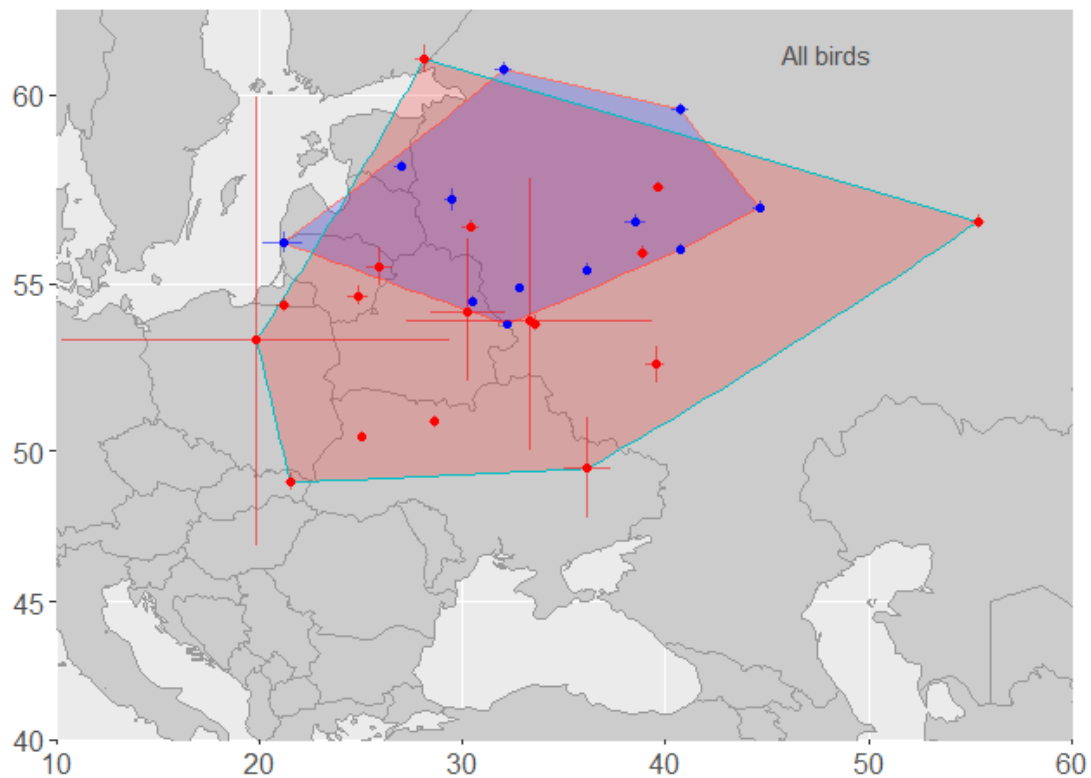
**Figure S9:** All birds summer unadjusted; all June locations (*SUMMER SEA JUNE* Table S1.a.4). Blue 2013, Red 2014;  $\pm 2$  standard errors in latitude and longitude are plotted for each location.

2013 0.603; 2014 2.049 million square km  
 593 ( $\pm 18$ ) vs 903 ( $\pm 32$ ) km ( $t_{28} = 6.4$ ,  $P = 0.0001$ )  
 Area diff = 1.446 mskm (99<sup>th</sup> percentile)  
 Random -0.567 to 1.147mskm  
 Overall 2.519 mskm = 24.7%  
 20.9% (99<sup>th</sup> percentile)  
 Random 35.3 to 75.2%



**Figure S10:** All birds summer adjusted; all June locations (*SUMMER SEA JUNE* Table S2.a.4). Blue 2013, Red 2014; +/- 2 standard errors in latitude and longitude are plotted for each location.

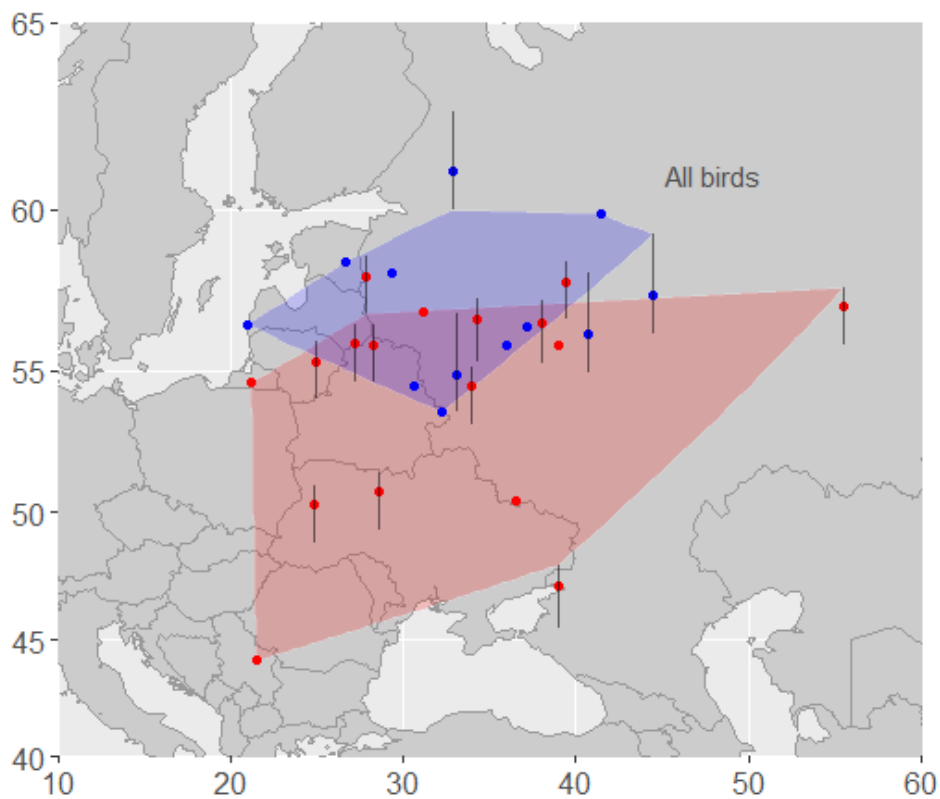
2013 0.603; 2014 1.814 million square km  
 593 (+/- 18) vs 824 (+/- 30) km ( $t_{28} = 4.7$ ,  $P = 0.0001$ )  
 Area diff = 1.211 mskm (99<sup>th</sup> percentile)  
 Random -0.454 to 0.897 mskm  
 Overall 1.835 mskm = 18.0%  
 31.1% (97<sup>th</sup> percentile)  
 Random 35.2 to 76.0%





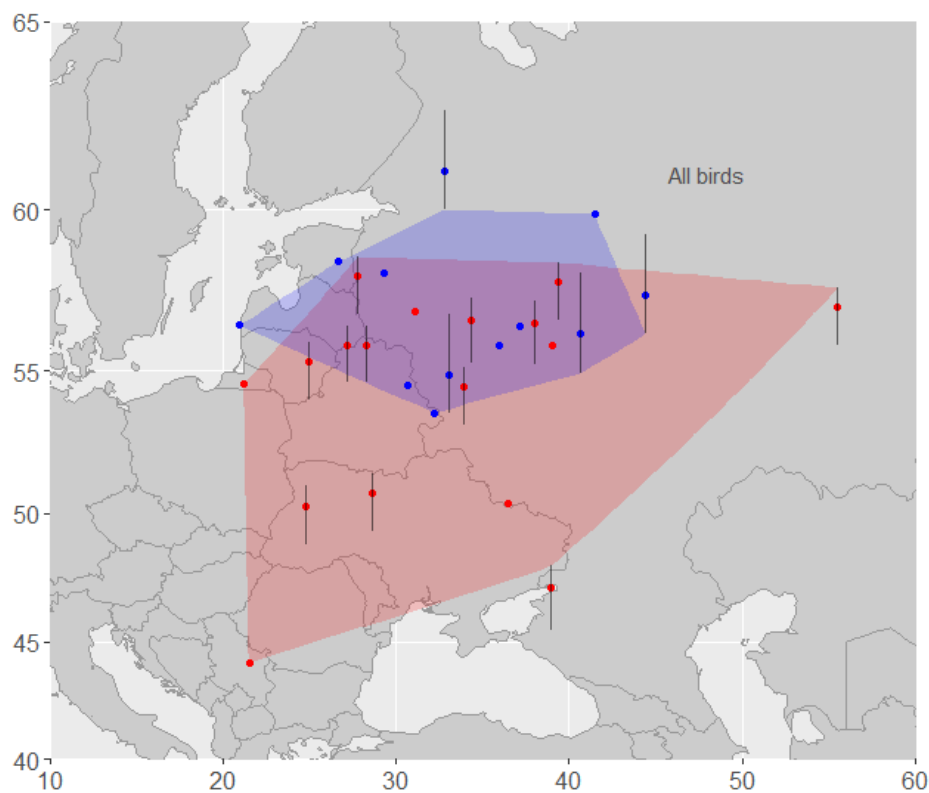
**Figure S11:** All birds summer SEA with SEA set to minimise total range for birds where H-E calibration did not give an individual value (*MINIMUM RANGE SUMMER SEA* Table S1.a.5). Blue 2013, Red 2014; +/- 2 standard errors in latitude and longitude are plotted for each location. The black vertical lines show the range of SEAs observed for all other loggers where the calibration was successful for that year.

2013 0.503; 2014 1.971 million square km  
 591 (+/- 19) vs 839 (+/- 33) km ( $t_{28} = 4.8$ ,  $P = 0.0001$ )  
 Area diff 1.468 mskm (99<sup>th</sup> percentile)  
 Random to -0.568 to 1.144 mskm  
 Overall 2.351 mskm = 23.1%  
 8.7% (99<sup>th</sup> percentile)  
 Random 32.1% to 75.9%



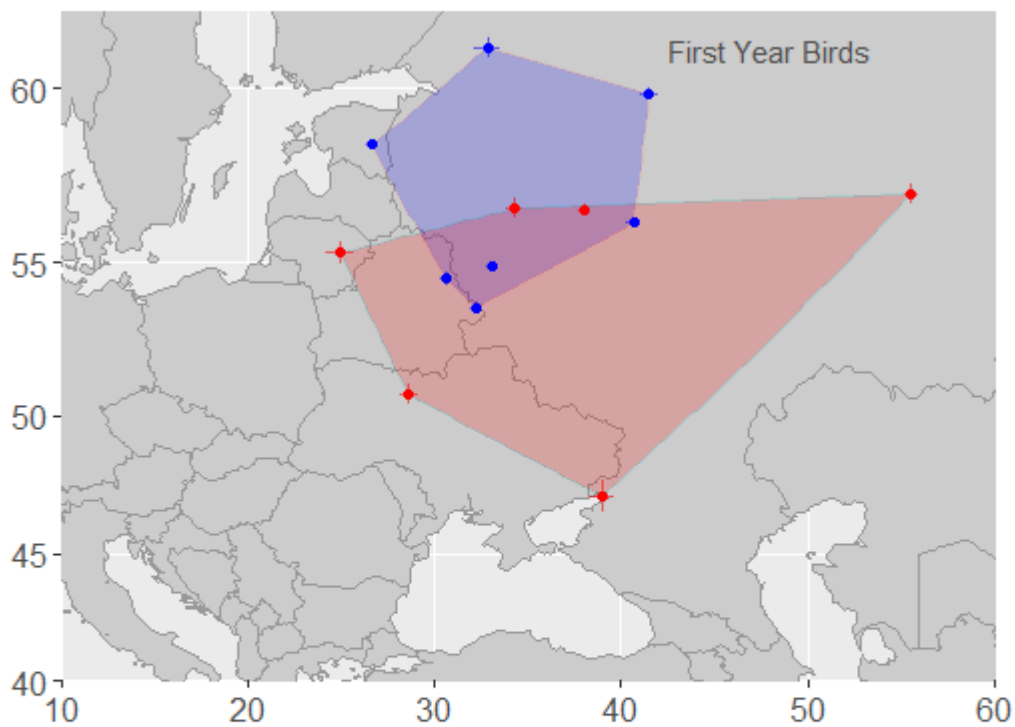
**Figure S12:** All birds summer SEA with SEA set at maximum to maximise range overlap for birds where H-E calibration did not give an individual value (*MAXIMUM OVERLAP SUMMER SEA* Table S1.a.6). Blue 2013, Red 2014;  $\pm 2$  standard errors in latitude and longitude are plotted for each location. The black vertical lines show the range of SEAs observed for all other loggers where the calibration was successful for that year.

2013 0.640; 2014 2.126 million square km  
 599 ( $\pm 18$ ) vs 870 ( $\pm 33$ ) km ( $t_{28} = 5.3$ ,  $P = 0.0001$ )  
 Area diff 1.486 mskm (99<sup>th</sup> percentile)  
 Random to -0.520 to 1.061 mskm  
 Overall 2.351 mskm = 23.1%  
 21.9% (99<sup>th</sup> percentile)  
 Random 34.0% to 76.6%



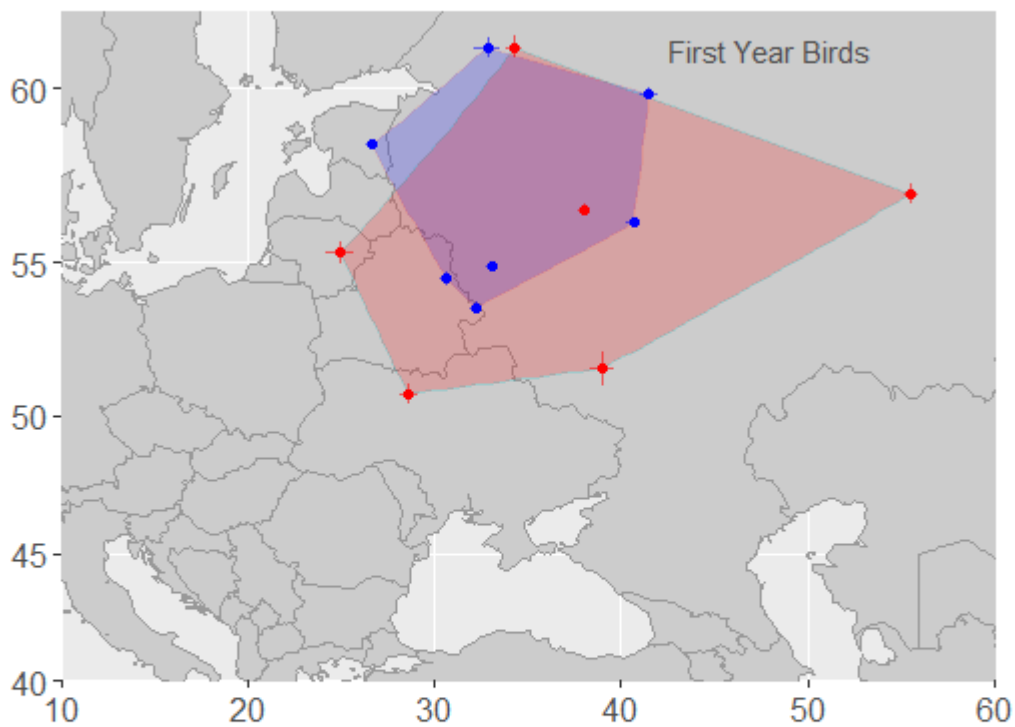
**Figure S13:** First year birds summer unadjusted (*BEST SUMMER SEA* Table S1.b.1). Blue 2013, Red 2014; +/- 2 standard errors in latitude and longitude are plotted for each location.

2013 0.470; 2014 1.199 million square km  
 598 (+/- 40) vs 1059 (+/- 72) km ( $t_{12} = 4.1$ ,  $P = 0.001$ )  
 Area diff = 0.729 mskm (98<sup>th</sup> percentile)  
 Random -0.857 to 0.589 mskm  
 Overall 1.700 mskm = 16.7%  
 7.9% (94<sup>th</sup> percentile)  
 Random 3.8% - 56.8%



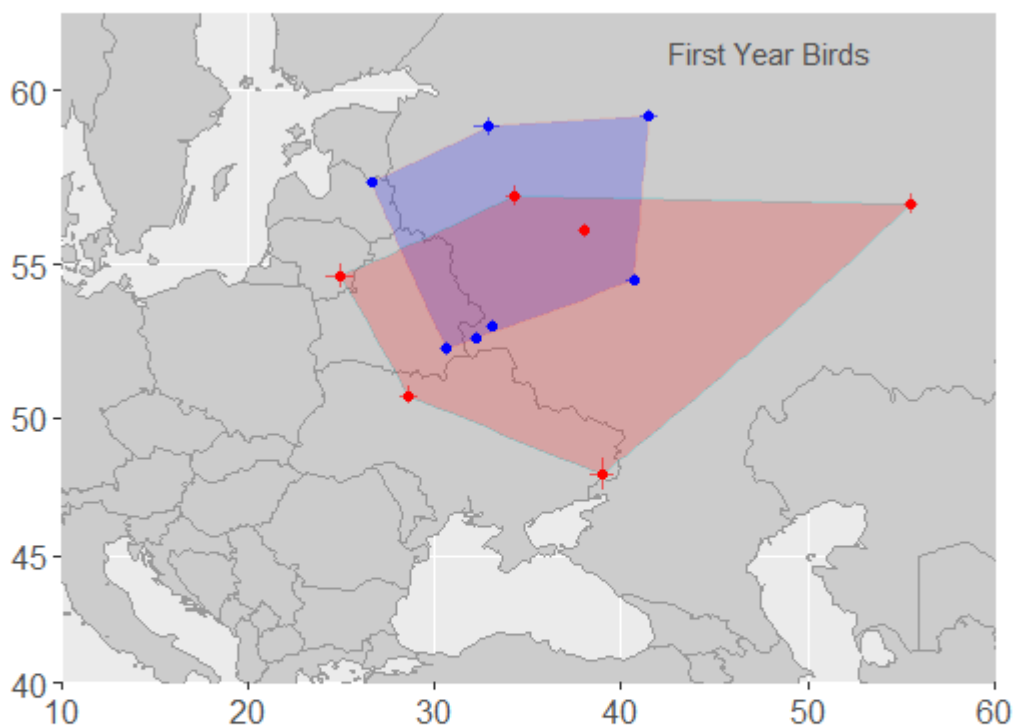
**Figure S14:** First year birds summer adjusted (*BEST SUMMER SEA* Table S2.b.1). Blue 2013, Red 2014;  
 +/- 2 standard errors in latitude and longitude are plotted for each location.

2013 0.470; 2014 1.232 million square km  
 598 (+/- 40) vs 1042 (+/- 63) km ( $t_{12} = 4.3$ ,  $P = 0.001$ )  
 Area diff = 0.762 mskm (99<sup>th</sup> percentile)  
 Random -0.651 to 0.486 mskm  
 Overall 1.275 mskm = 13.3%  
 33.2% (49<sup>th</sup> percentile)  
 Random 4.2% - 60.3%



899 **Figure S15:** First year birds winter unadjusted (*WINTER SEA* Table S1.b.2). Blue 2013, Red 2014; +/- 2  
900 standard errors in latitude and longitude are plotted for each location.

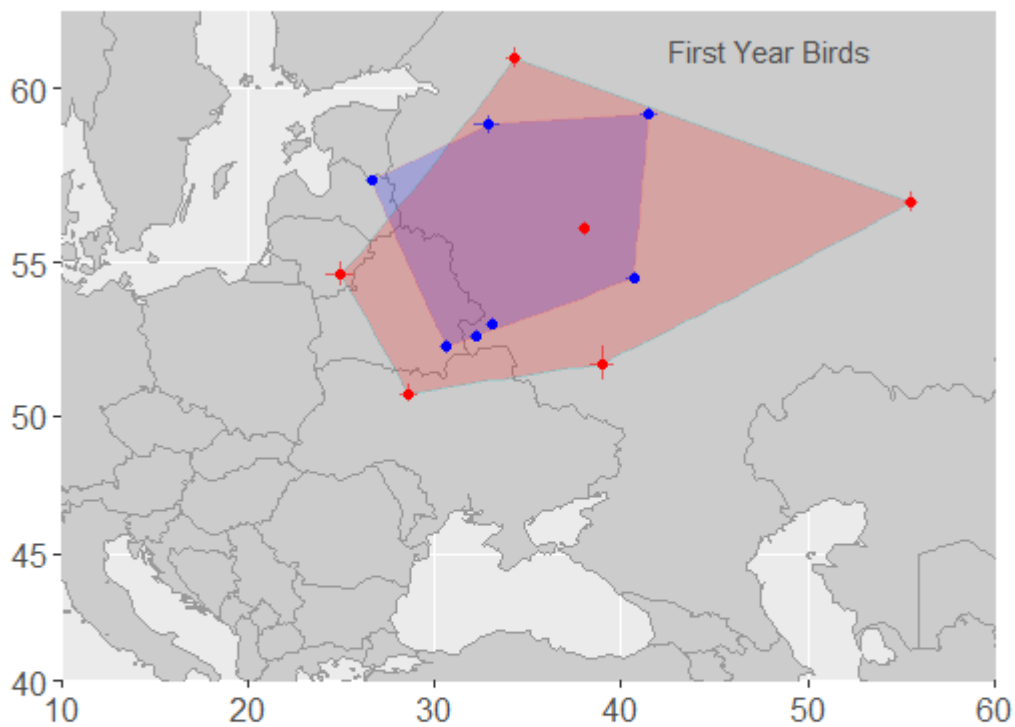
901 2013 0.487; 2014 1.145 million square km  
902 620 (+/- 43) vs 1032 (+/- 73) km ( $t_{12} = 3.6$ ,  $P = 0.003$ )  
903 Area diff 0.658 mskm (96<sup>th</sup> percentile)  
904 Random -0.706 to 0.590  
905 Overall 1.462 mskm = 14.3%  
906 18.9% (76<sup>th</sup> percentile)  
907 Random 4.8% - 58.6%



908

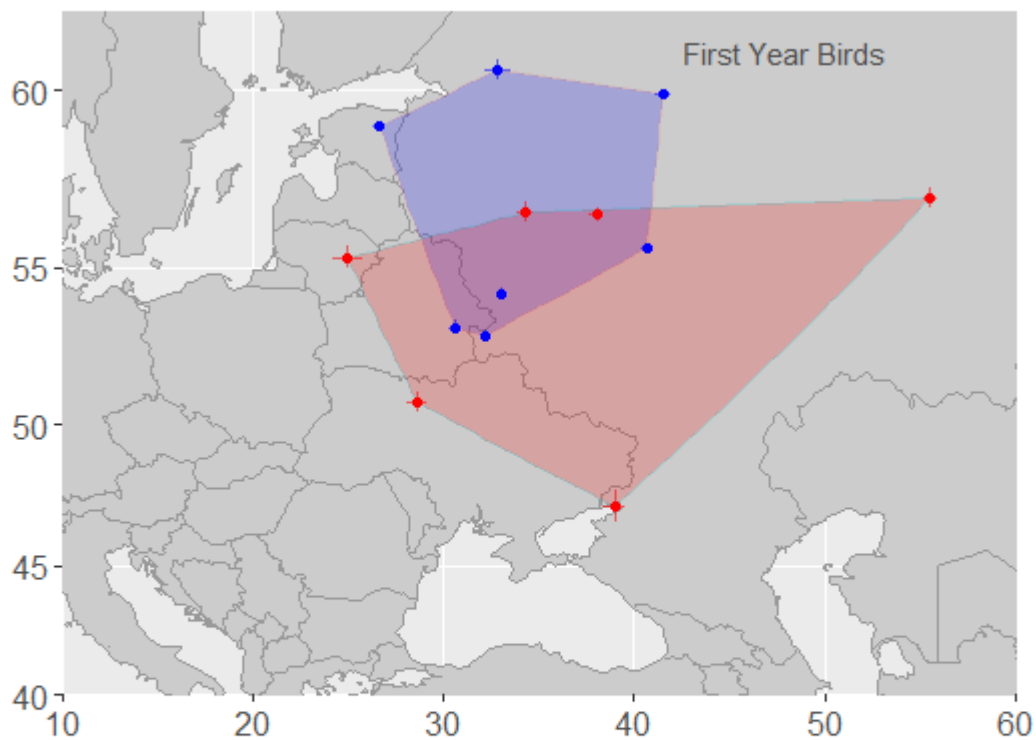
**Figure S16:** First year birds winter adjusted (*WINTER SEA* Table S2.b.2). Blue 2013, Red 2014; +/- 2 standard errors in latitude and longitude are plotted for each location.

2013 0.487; 2014 1.176 million square km  
 620 (+/- 43) vs 1028 (+/- 66) km ( $t_{12} = 3.8$ ,  $P = 0.002$ )  
 Area diff 0.689 mskm (99<sup>th</sup> percentile)  
 Random to -0.605 to 0.442  
 Overall 1.220 mskm = 12.0%  
 39.4% (29<sup>th</sup> percentile)  
 Random 3.6% - 58.8%

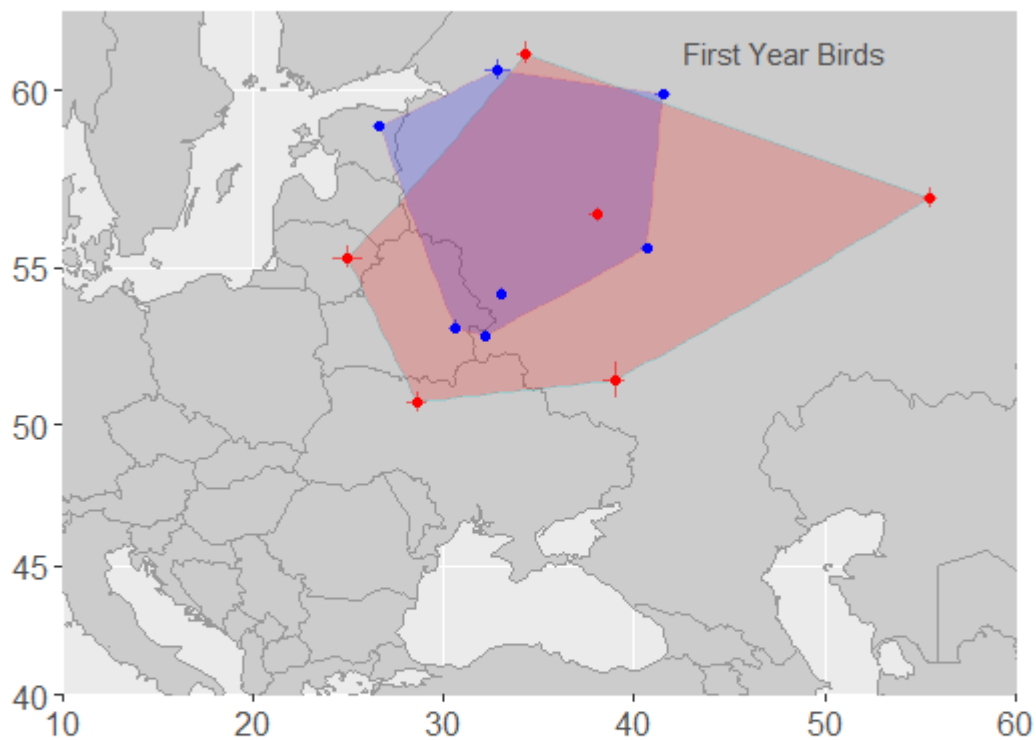


**Figure S17:** First year birds average unadjusted (*AVERAGE SEA* Table S1.b.3). Blue 2013, Red 2014;  
 +/- 2 standard errors in latitude and longitude are plotted for each location.

2013 0.518; 2014 1.199 million square km  
 640 (+/- 41) vs 1060 (+/- 72) km ( $t_{12} = 3.7$ ,  $P = 0.003$ )  
 Area diff 0.681 mskm (95<sup>th</sup> percentile)  
 Random -0.832 to 0.622  
 Overall 1.689 mskm = 16.6%  
 11.6% (90<sup>th</sup> percentile)  
 Random 3.9% - 60.2%



929 **Figure S18:** First year birds average adjusted (*AVERAGE SEA* Table S2.b.3). Blue 2013, Red 2014; +/-  
 930 2 standard errors in latitude and longitude are plotted for each location.  
 931 2013 0.518; 2014 1.227 million square km  
 932 640 (+/- 41) vs 1040 (+/- 63) km ( $t_{12} = 3.8$ ,  $P = 0.002$ )  
 933 Area diff 0.709 mskm (99<sup>th</sup> percentile)  
 934 Random -0.660 to 0.449  
 935 Overall 1.327 mskm = 10.0%  
 936 33.4% (48<sup>th</sup> percentile)  
 937 Random 2.2% - 60.1%



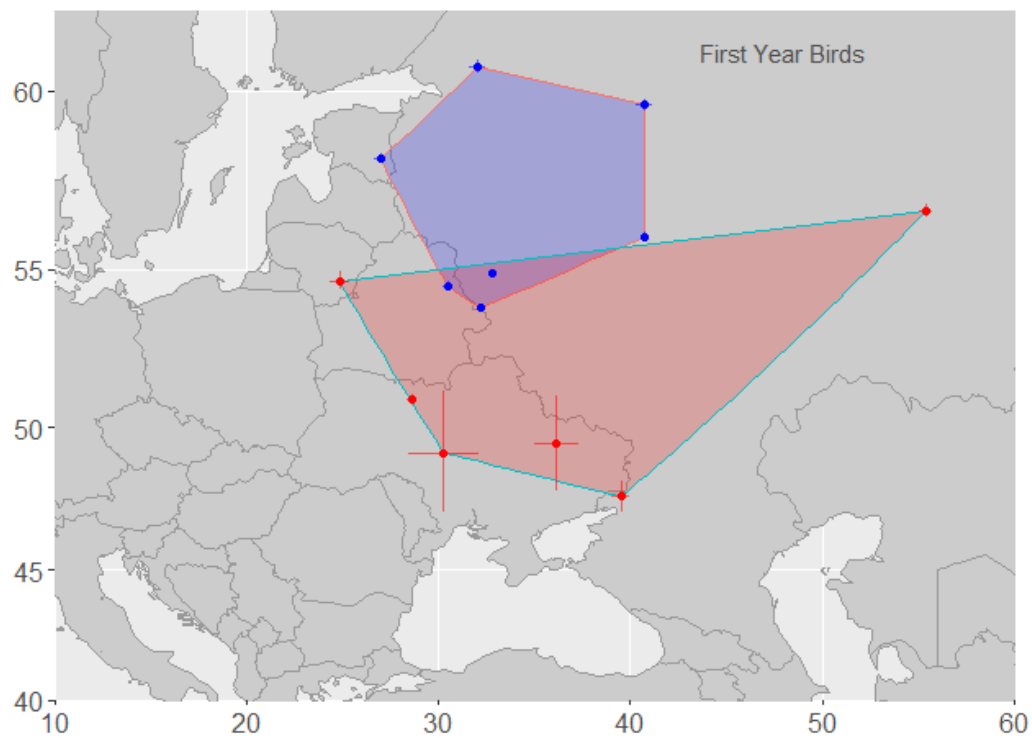
938



**Figure S19:** First year birds summer unadjusted; all June locations (*SUMMER SEA JUNE* Table S1.b.4).

Blue 2013, Red 2014;  $\pm 2$  standard errors in latitude and longitude are plotted for each location.

2013 0.429; 2014 1.153 million square km  
567 ( $\pm 40$ ) vs 1013 ( $\pm 70$ ) km ( $t_{12} = 3.3$ ,  $P = 0.007$ )  
Area diff 0.681 mskm (97<sup>th</sup> percentile)  
Random -0.780 to 0.577  
Overall 1.672 mskm = 16.4%  
2.7% (96<sup>th</sup> percentile)  
Random 4.6% to 59.3%



948  
949

**Figure S20:** First year birds summer adjusted; all June locations (*SUMMER SEA JUNE* Table 2.b.4). Blue

2013, Red 2014;  $\pm 2$  standard errors in latitude and longitude are plotted for each location.

2013 0.429; 2014 0.829 million square km

567 ( $\pm 40$ ) vs 950 ( $\pm 65$ ) km ( $t_{12} = 3.1$ ,  $P = 0.01$ )

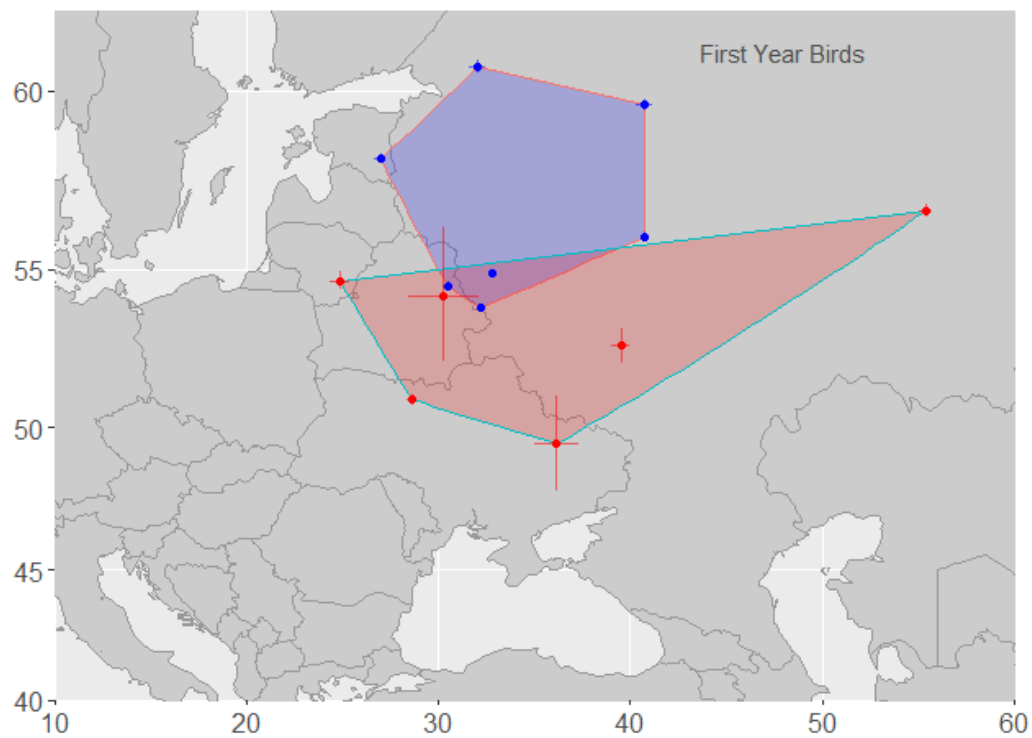
Area diff 0.400 mskm (88<sup>th</sup> percentile)

Random -0.652 to 0.521

Overall 1.348 mskm = 13.2%

3.3% (96<sup>th</sup> percentile)

Random 4.1% to 58.5%



959  
960